

Unifying Advective and Diffusive Descriptions of Bedform Pumping in the Benthic Biolayer of Streams

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- Parallel advective and diffusive models of solute transport through the benthic biolayer of streams by bedform pumping are derived
- The two models are functionally equivalent provided that the diffusion model's dispersion coefficient decays exponentially with depth
- Both frameworks closely reproduce measured patterns and rates of bedform pumping, and provide complementary predictive capabilities

Abstract. Many water quality and ecosystem functions performed by streams occur in the benthic biolayer, the biologically active upper (~ 5 cm) layer of the streambed. Solute transport through the benthic biolayer is facilitated by bedform pumping, a physical process in which dynamic and static pressure variations over the surface of stationary bedforms (e.g., ripples and dunes) drive flow across the sediment-water interface. In this paper we derive two predictive modeling frameworks, one advective and the other diffusive, for solute transport through the benthic biolayer by bedform pumping. Both frameworks closely reproduce patterns and rates of bedform pumping previously measured in the laboratory, provided that the diffusion model's dispersion coefficient declines exponentially with depth. They are also functionally equivalent, such that parameter sets inferred from the advective model can be applied to the diffusive model, and vice versa. The functional equivalence and complementary strengths of these two models expands the range of questions that can be answered, for example by adopting the advective model to study the effects of geomorphic processes (such as bedform adjustments to land use change) on flow-dependent processes, and the diffusive model to study problems where multiple transport mechanisms combine (such as bedform pumping and turbulent diffusion). By unifying advective and diffusive descriptions of bedform pumping, our analytical results provide a straightforward and computationally efficient approach for predicting, and better understanding, solute transport in the benthic biolayer of streams and coastal sediments.

46 **Plain Language Summary**

47 How far and fast pollutants travel downstream is often conditioned on what happens in a thin
48 veneer of biologically active bottom sediments called the benthic biolayer. However, before a
49 pollutant can be removed in the benthic biolayer it must first be transported across the sediment-
50 water interface and through the interstitial fluids of these surficial sediments. In this paper we
51 demonstrate that one important mechanism for transporting solutes to, and through, the benthic
52 biolayer—bedform pumping—can be interchangeably represented as either a two-dimensional
53 advective process or a one-dimensional dispersion process. The complementary nature of these
54 models expands the range of benthic biolayer processes that can be studied and predicted with
55 the end goal of improving coastal and stream water quality.

56

1. Introduction

The movement of water into and out of the hyporheic zone, or “hyporheic exchange”, occurs over a wide range of spatial (and temporal) scales, from >10 km (>1 year) to <1 m (<1 hr) (Boano et al., 2014; Gomez-Velez and Harvey, 2014; Wörman et al., 2007). This $>10^3$ range of temporal and spatial scales raises trade-offs—relative to residence times, reaction times, and exchange rates—that can influence the hyporheic zone’s ability to process nutrients and other pollutants (Harvey et al., 2013). For example, Gomez-Velez et al. (2015) evaluated the residence time/exchange rate trade-off for aerobic respiration and denitrification in the Mississippi River Network, calculating for each reach a so-called Reaction Significance Factor, RSF (Harvey et al., 2013). In the RSF framework, more nutrients are removed when hyporheic zone residence times are comparable to reaction times and the uptake length is short compared to the reach length (i.e., the RSF is large). These authors found that the smallest scales of hyporheic exchange are the most important for nutrient processing in streams, with RSFs consistently larger for vertical exchange over submerged ripples and dunes (length-scales of the order of 10^0 m) compared to lateral exchange over larger geomorphic features such as river bars and meandering banks (length-scales of the order of 10^2 to 10^3 m). This conclusion, which is based on physical arguments, is reinforced by findings that microbial biomass and nitrification and denitrification potential tend to be concentrated in the upper 5 cm of the streambed, a region of the hyporheic zone known as the “benthic biolayer” (Tomasek et al., 2018; Knapp et al., 2017; Caruso et al., 2017). Collectively, these results underscore the importance of elucidating physical mechanisms responsible for hyporheic exchange at the scale where nutrient transformations primarily occur; that is, in the benthic biolayer.

At the scale of the benthic biolayer one important driver of hyporheic exchange is bedform pumping, which occurs when dynamic and static pressure variations over the surface of bedforms (e.g., ripples and dunes) drive flow across the sediment-water interface (SWI) in spatially isolated upwelling and downwelling zones (Azizian et al., 2015; Azizian et al., 2017; Grant et al., 2012; Grant et al., 2014; Fleckenstein et al., 2010; Cardenas et al., 2008; Elliot and Brooks, 1997a,b; Thibodeaux and Boyle, 1987) (**Figure 1a**). Since its discovery in 1987 (Thibodeaux and Boyle, 1987), a number of analytical models have been proposed to describe bedform pumping and its influence on stream water quality (reviewed in Boano et al., 2014). Generally, these models can be grouped depending on whether they conceptualize bedform pumping as an advective or diffusive process. Advective models are notable for their relatively faithful representation of the laminar flow fields generated by bedform pumping (Elliott and Brooks 1997a,b). An advantage of diffusive models is their ability to incorporate multiple mechanisms for mass transport across the SWI (i.e., not just bedform pumping) including molecular diffusion, turbulent diffusion and dispersion (Voermans et al., 2017; Voermans et al., 2018; Grant et al., 2018a; Grant et al., 2018b; Grant et al., in review).

As commonly implemented, both types of analytical models rely on multiple assumptions that limit their practical utility: (1) solute concentration in the overlying water column is assumed constant in time; (2) two-way coupling across the SWI—whereby mass transfer out of the streambed alters mass concentration in the overlying water column which, in turn, alters mass transfer into the streambed, and so on—is not accounted for; (3) diffusive mixing in the streambed is constant in depth, while the interstitial flow field generated by bedform pumping

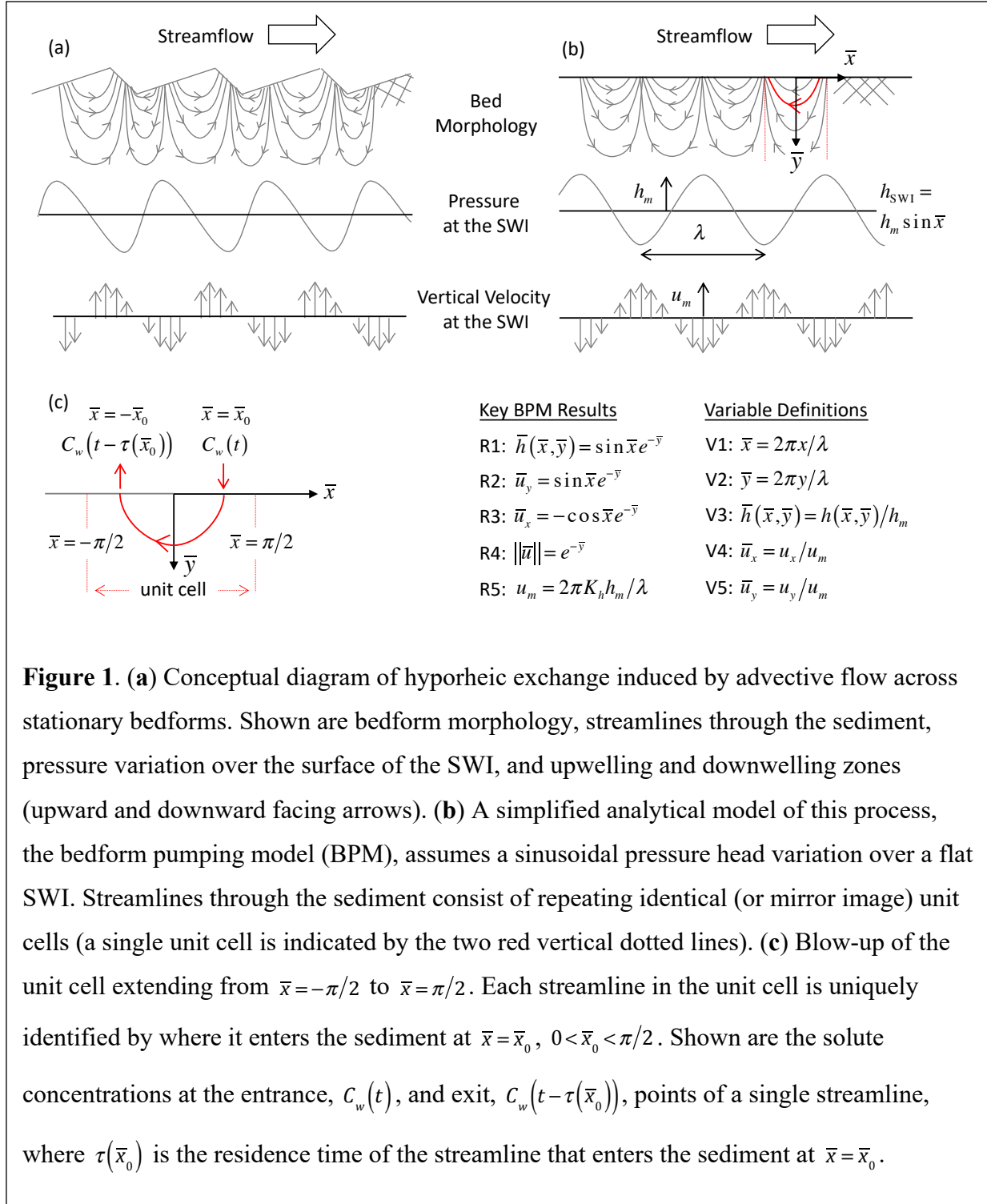


Figure 1. (a) Conceptual diagram of hyporheic exchange induced by advective flow across stationary bedforms. Shown are bedform morphology, streamlines through the sediment, pressure variation over the surface of the SWI, and upwelling and downwelling zones (upward and downward facing arrows). (b) A simplified analytical model of this process, the bedform pumping model (BPM), assumes a sinusoidal pressure head variation over a flat SWI. Streamlines through the sediment consist of repeating identical (or mirror image) unit cells (a single unit cell is indicated by the two red vertical dotted lines). (c) Blow-up of the unit cell extending from $\bar{x} = -\pi/2$ to $\bar{x} = \pi/2$. Each streamline in the unit cell is uniquely identified by where it enters the sediment at $\bar{x} = \bar{x}_0$, $0 < \bar{x}_0 < \pi/2$. Shown are the solute concentrations at the entrance, $C_w(t)$, and exit, $C_w(t - \tau(\bar{x}_0))$, points of a single streamline, where $\tau(\bar{x}_0)$ is the residence time of the streamline that enters the sediment at $\bar{x} = \bar{x}_0$.

decays exponentially (Elliott and Brooks, 1997a); and (4) some published diffusive models fail to account for the finite porosity of the streambed, a violation of mass balance that can bias estimates of the diffusivity downwards by a factor of ten (Grant et al., 2012).

In this paper we derive two parallel analytical frameworks, one advective and the other diffusive, that collectively address the model limitations noted above. The paper is organized as follows. In **Section 2** we review a canonical analytical model for advective bedform pumping originally developed by Elliott and Brooks (1997a,b) (**Section 2.1**), show that its residence time distribution closely follows the extreme value Fréchet distribution (**Section 2.2**), and derive from this result a set of fully coupled solutions for the evolution of solute concentration in the water and sediment columns of a closed system (**Section 2.3**). In **Section 3** we derive a parallel diffusive analytical framework for bedform pumping (**Section 3.1**), show how the choice of a diffusivity profile (constant or exponentially declining) leads to different Green's function (Leij et al., 2000) representations of mass transport in the streambed (**Section 3.2**), and then derive from these Green's functions a set of fully coupled solutions for the evolution of solute concentration in the water and sediment columns of a closed system (**Section 3.3**). We test these models against previously published measurements of unsteady solute transport across artificial and natural bedforms in a recirculating flume (**Section 4**). Discussion of these results are presented in **Section 5** and conclusions in **Section 6**.

2. Advective Bedform Pumping Model (BPM)

2.1 Canonical Solution by Elliott and Brooks (EB)

A canonical advective model of bedform pumping (originally solved by Vaux (1968) and expanded on by Elliott and Brooks (1997a,b)), hereafter referred to as the bedform pumping model (**BPM**), assumes that hyporheic exchange is driven by a sinusoidal variation of pressure head over a flat SWI (**Figure 1b**). The wavelength λ [L] of the pressure wave corresponds to the wavelength of the bedform, and the trough and peak of the pressure wave correspond to where the velocity boundary layer detaches (at the bedform crest) and reattaches (on the lee side of the

bedform), respectively (Cardenas and Wilson, 2007a,b; Sawyer and Cardenas, 2009). If the hydraulic conductivity K_h [$L T^{-1}$] and porosity θ of the streambed are constant, Darcy's Law and the continuity equation can be jointly solved to yield the BPM's well-known formulae for the two-dimensional pressure head distribution and velocity field in the interstitial pores of the hyporheic zone (equations (R1) - (R5), **Figure 1b**) (Elliot and Brooks 1997a,b).

As documented in Supplemental Information (**Text S1**), if the sediment bed is initially solute free (at $\bar{t}=0$) and the solute in question is conservative (i.e., inert and does not adsorb to sediments) the average interfacial flux, $J(\bar{t})$ [$M L^{-2} T^{-1}$], of mass into the streambed can be represented as a convolution over all past water column concentrations, $C_w(\bar{t})$ [$M L^{-3}$]:

$$J(\bar{t}) = \frac{u_m}{\pi} \left[C_w(\bar{t}) - \int_0^{\bar{t}} C_w(\bar{t} - \bar{\tau}) f_{RTD}(\bar{\tau}) d\bar{\tau} \right] \quad (1a)$$

$$f_{RTD}(\bar{\tau}) = \frac{\sin[\bar{x}_0(\bar{\tau})] \cos[\bar{x}_0(\bar{\tau})]}{1 + \bar{x}_0(\bar{\tau}) \tan[\bar{x}_0(\bar{\tau})]} \quad (1b)$$

$$\bar{\tau} = \frac{\bar{x}_0(\bar{\tau})}{\cos[\bar{x}_0(\bar{\tau})]} \quad (1c)$$

The function $f_{RTD}(\bar{\tau})$ [-] is the probability density function (PDF) form of the BPM's residence time distribution (RTD), defined such that the quantity $f_{RTD}(\bar{\tau}) d\bar{\tau}$ is the fraction of water circulating through the hyporheic zone with dimensionless residence times in the range $\bar{\tau}$ to $\bar{\tau} + d\bar{\tau}$. The variable u_m appearing on the right hand side of equation (1a) is the maximum Darcy flux of water across the SWI, and time and residence time ($\bar{t} = t/t_T$ and $\bar{\tau} = \tau/t_T$, respectively) have been scaled by a characteristic timescale for the transport of solute through a bedform: $t_T = \lambda \theta / \pi u_m$ (all BPM variables defined in **Figure 1**).

2.2. The Fréchet Distribution and the BPM's Residence Time Distribution (RTD)

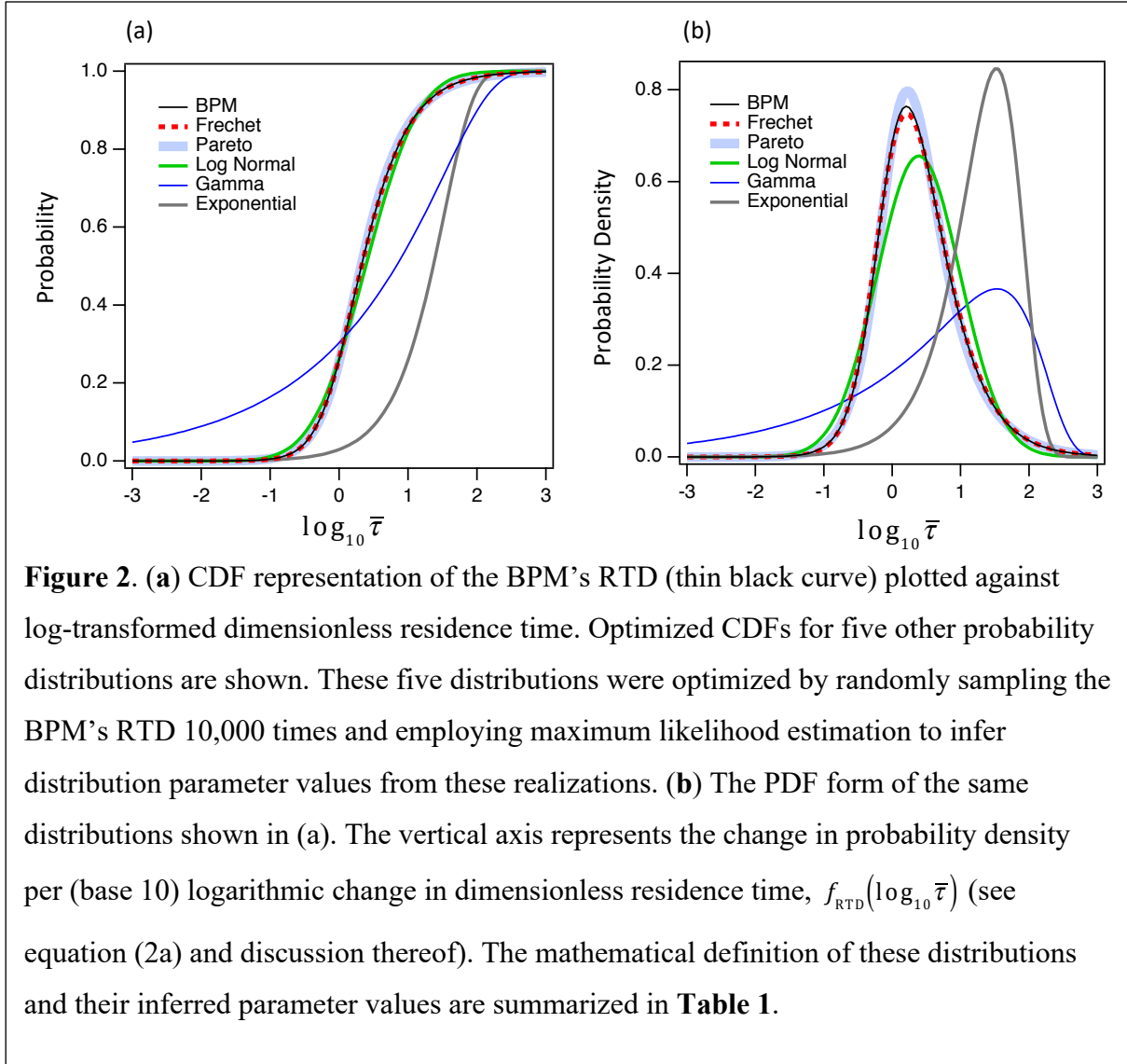
For any choice of the dimensionless residence time $\bar{\tau}$, numerical evaluation of the BPM's RTD requires two steps. First, the dimensionless starting position, $\bar{x}_0(\bar{\tau})$ [-], of the streamline in the unit cell with dimensionless time $\bar{\tau}$ (see **Figure 1c**) is obtained by numerically solving the implicit expression for $\bar{x}_0(\bar{\tau})$ (equation (1c)). This estimate of $\bar{x}_0(\bar{\tau})$ is then substituted into the RTD formula (equation (1b)) to obtain the fraction of flow leaving the hyporheic zone with that dimensionless residence time. Because hyporheic zone residence times vary over many orders of magnitude, it is convenient to divide the unit area under the RTD into evenly spaced logarithmic increments of dimensionless residence time (Azizian et al., 2017):

$$f_{\text{RTD}}(\log_{10} \bar{\tau}) = \frac{dF_{\text{RTD}}}{d\log_{10} \bar{\tau}} = 2.303 \bar{\tau} f_{\text{RTD}}(\bar{\tau}) \quad (2a)$$

$$F_{\text{RTD}}(\bar{\tau}) = 1 - \cos[\bar{x}_0(\bar{\tau})] \quad (2b)$$

The cumulative distribution function (CDF) form of the RTD appearing in equation (2b), $F_{\text{RTD}}(\bar{\tau})$ [-], is defined as the fraction of water circulating through the hyporheic zone with dimensionless residence time of $\bar{\tau}$ or younger; the PDF and CDF forms of the RTD are related in the usual way: $f_{\text{RTD}}(\bar{\tau}) = dF_{\text{RTD}}(\bar{\tau})/d\bar{\tau}$. As demonstrated in the Supplemental Information (**Text S2**), our definition of $F_{\text{RTD}}(\bar{\tau})$ is mathematically consistent with the one derived by Elliott and Brooks (hereafter, **EB**) in their original publication of the BPM (Elliott and Brooks, 1997a).

The BPM's RTD spans a thousand-fold change in dimensionless residence times, from $\bar{\tau} < 0.1$ to $\bar{\tau} > 100$ (black curves in **Figures 2a** and **2b**). It is well described by both the Fréchet and Pareto distributions, reasonably well described by the Log-Normal distribution, and poorly described by the Gamma and Exponential distributions (colored curves in the figure). The

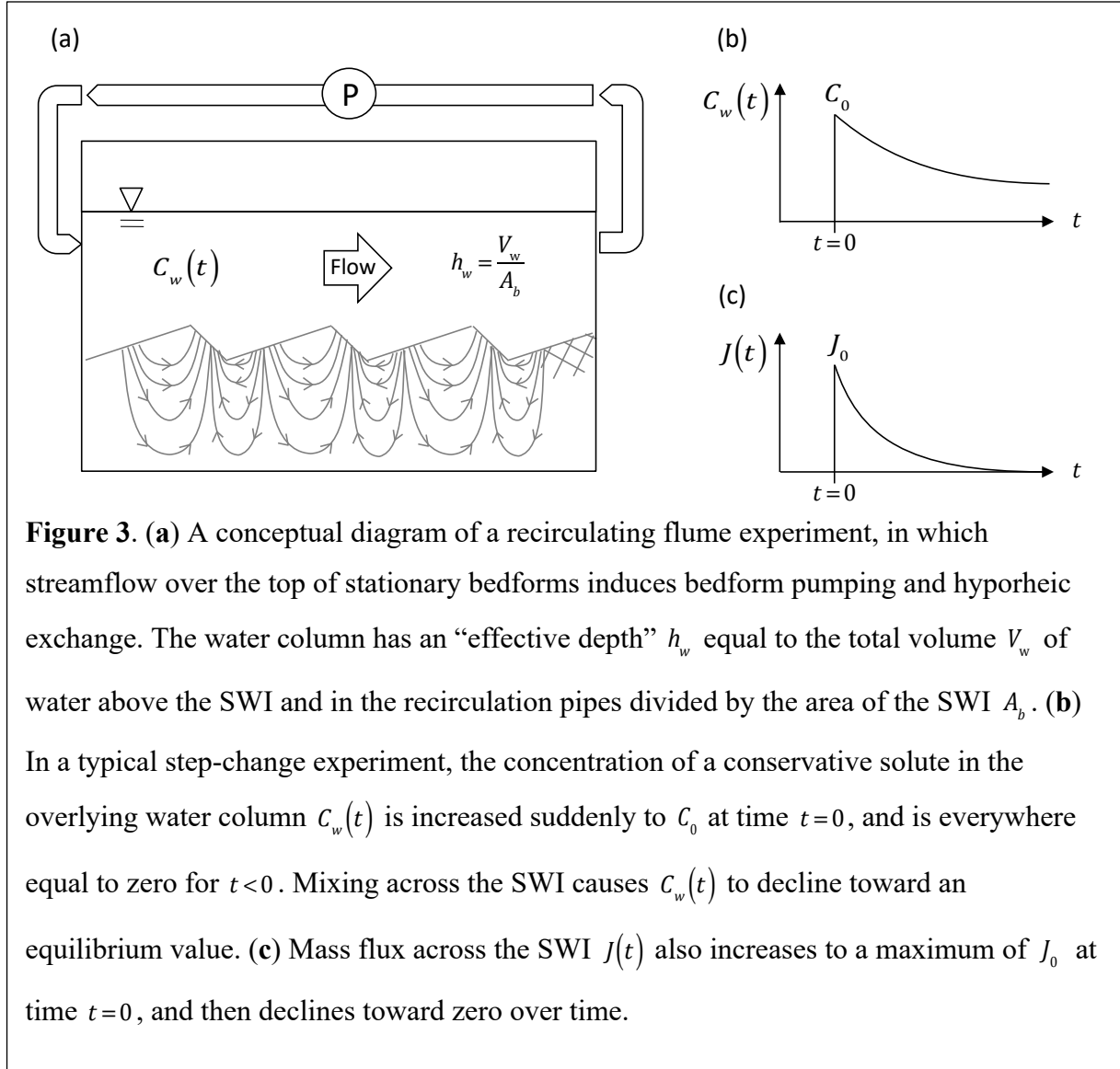


166 remarkable similarity between the BPM's RTD and the Fréchet distribution—a heavy-tailed
 167 extreme value distribution (Kotz and Nadarajah, 2000)—has not, to our knowledge, been noted
 168 in the literature. More commonly, the power-law or Pareto distribution is adopted to represent
 169 hyporheic exchange (Bottacin-Busolin and Marion, 2010). However, the three-parameter version
 170 of the Pareto distribution was required to obtain a reasonable match to the BPM's RTD and, even
 171 then, the Pareto distribution ranked second behind the (two-parameter) Fréchet distribution (see
 172 Kolmogorov-Smirnov ranking in **Table 1**). The log-normal distribution, which is sometimes
 173 used to model residence times in the hyporheic zone (e.g., Wörman et al., 2002; Azizian et al.,

2017), underpredicts the RTD's heavy tail but is otherwise comparable to the BPM's RTD (compare green and black curves, **Figure 2b**). The Gamma distribution has been used to represent the RTD of water parcels moving through hillslopes (Kirchner et al., 2000; Leray et al., 2016) while the Exponential RTD underpins the Transient Storage Model, a popular hyporheic exchange modeling framework (Knapp and Kelleher, 2020)). Based on the results in **Figure 2** these last two distributions should not be used to represent the BPM's RTD. For the analysis that follows we adopted the optimized Fréchet distribution in place of the BPM's RTD for three reasons: (1) the Fréchet distribution is parsimonious and closely matches the BPM's actual RTD (**Table 1** and **Figure 2**); (2) this approach side steps the numerical challenges associated with the two-step process required to solve the BPM's RTD (see equation (1b) and discussion thereof); and, (3) the Laplace Transform of the Fréchet distribution can be computed analytically, which simplifies the mass balance analysis described next.

2.3 Unsteady Solute Concentration in the Water Column of a Closed System

An example of bedform pumping in a closed system is the recirculating flume set-up illustrated in **Figure 3a**. A mass M of a conservative solute is added to the water column of a solute-free recirculating flume at time $t=0$. After a short mixing period, the concentration in the water column is approximately $C_0 = M/V_w$ where V_w is the volume of water above the sediment bed and in the recirculating pipes. At this point in time, the second term on the right hand side of equation (1a) is negligible (because no solute has yet passed through the hyporheic zone and returned to the stream) and therefore the BPM predicts that the initial flux of solute into the bed should be: $J_0 = C_0 u_m / \pi$. With increasing elapsed time ($t > 0$) the solute concentration in the overlying water column declines (**Figure 3b**), the integral term in equation (1a) becomes progressively larger in magnitude (as solute in the streambed begins to return to the stream), and the net flux across the



197 SWI asymptotically approaches zero (**Figure 3c**). In practice, if the experiment runs long
 198 enough, the water column solute concentration will approach an equilibrium value, C_{eq} [M L⁻³],
 199 reflecting a well-mixed final state in which the solute concentration is the same everywhere in
 200 the overlying water column and the interstitial fluids of the sediment bed:

$$201 \quad \bar{C}_{eq} = C_{eq} / C_0 = 1 / (d_b \theta / h_w + 1) \quad (3)$$

New variables in equation (3) include the effective water depth h_w [L] (equal to the volume of water V_w [M L⁻³] in the overlying water column and recirculating pipes divided by the surface area A_b [L²] of the bed, $h_w = V_w / A_b$), sediment bed depth d_b [L] and porosity θ [-]. As the BPM assumes the streambed is infinitely deep, this analytical model never achieves equilibrium.

One challenge associated with deriving a solution for the recirculating flume problem illustrated in **Figure 3a** is the two-way coupling of solute concentrations above and below the SWI. This two-way coupling is evident when mass balance is performed over the recirculating flume's water column:

$$\frac{T}{t_T} \frac{d\bar{C}_w}{d\bar{t}} = -\bar{C}_w(\bar{t}) + \int_0^{\bar{t}} \bar{C}_w(\bar{t} - \bar{\tau}) f_{\text{RTD}}(\bar{\tau}) d\bar{\tau} \quad (4)$$

Here, the water column concentration has been scaled by its initial concentration at $t=0$ ($\bar{C} = C/C_0$) and the variable $T = h_w \pi / u_m$ represents a characteristic timescale for all water in the overlying water column and recirculating pipes to undergo hyporheic exchange. Two-way coupling manifests mathematically as a dependence of the time rate of change of the water column solute concentration (left hand side of equation (4)) on the entire past history of water column solute concentration filtered through the hyporheic zone's RTD (convolution integral on the right hand side of equation (4)). In addition to providing an elegant interpretation of two-way coupling, the convolution representation of hyporheic exchange flux permits an analytical solution to the overall mass balance problem. This is because the Laplace Transform of a convolution of two variables is equal to the product of their respective Laplace Transforms (Graff, 2004). Thus, after applying the Laplace Transform to equation (4), solving for the solute concentration in the water column becomes a simple algebraic exercise:

$$\bar{C}_w(\bar{t}) = \mathcal{L}^{-1} \left[\frac{\bar{T}}{\bar{s}\bar{T} + 1 - \tilde{f}_{\text{RTD}}(\bar{s}; \beta, \mu)} \right] \quad (5a)$$

Here, the variable $\bar{T} = T/t_r$ is a dimensionless timescale for hyporheic zone processing of water above the streambed, $\bar{s} = s t_r$ is a dimensionless form of the Laplace transform variable and the symbol $\mathcal{L}^{-1}[\cdot]$ denotes the inverse Laplace Transform which, in practice, is solved numerically (see **Section 4.1**). The Laplace Transform of the Fréchet distribution can be computed analytically by applying the Right Shift rule (Graf, 2004) where $K_1(\cdot)$ is the modified Bessel function of the second kind and u is a dummy integration variable:

$$\tilde{f}_{\text{RTD}}(\bar{s}; \beta, \mu) = \frac{e^{\mu \bar{s}}}{1 - e^{-\beta/\mu}} \left(2\sqrt{\beta \bar{s}} K_1(2\sqrt{\beta \bar{s}}) - \beta \int_0^\mu e^{-\bar{s}u} \frac{e^{-\beta/u}}{u^2} du \right) \quad (5b)$$

Because the Fréchet distribution parameters (β , μ) are known (see **Table 1**), we can infer from equation (5a) that $\bar{C}_w(\bar{t})$ depends on a single dimensionless parameter, $\bar{T} = T/t_r$. The two timescales appearing in this dimensionless parameter depend on physical characteristics of the recirculating flume as follows:

$$t_r = \frac{\lambda^2 \theta}{2\pi^2 K_h h_m} \quad (6a)$$

$$T = \frac{h_w \lambda}{2K_h h_m} \quad (6b)$$

Therefore, implementation of this analytical solution requires knowledge of the bedform wavelength λ , streambed porosity θ , streambed hydraulic conductivity K_h , the effective depth of the water column (taking into account the water in the recirculating system for the system illustrated in **Figure 3a**) h_w , and the half-amplitude of the pressure head variation h_m . With the

exception of h_m , these parameters are readily measured or predicted (e.g., the hydraulic conductivity can be estimated from the median grain diameter of unconsolidated sediments, for example using the Kozeny-Carman equation (McCabe et al. 2010)). To estimate h_m , the widely cited empirical formula proposed by EB (based on pressure measurements over a triangular bedform reported by Fehlman (1985)) can be employed:

$$h_m = 0.28 \frac{V^2}{2g} \left(\frac{H/d}{0.34} \right)^\gamma \quad (7)$$

The variables V [$L T^{-1}$] and d [L] represent, respectively, the average velocity and depth of the overlying stream, H [L] is bedform height, and the empirical exponent is taken as either $\gamma = 3/8$ (if $H/d < 0.34$) or $\gamma = 3/2$ (if $H/d \geq 0.34$). The value of the multiplicative constant (0.28) on the right hand side of equation (7) can be adjusted depending on the height-to-wavelength ratio of the bedform (Shen et al., 1990; Fox et al., 2014).

2.4 Unsteady Interstitial Solute Concentration in the Sediment Column of a Closed System

A corresponding analytical solution can be derived for the spatiotemporal evolution of solute concentration in the interstitial fluids of the sediment bed. The solution is premised on the idea that the interstitial concentration of a conservative solute at any dimensionless time \bar{t} is equal to the concentration that was present in the water column some location-dependent dimensionless residence time ago: $\bar{C}_s(\bar{x}, \bar{y}, \bar{t}) = \bar{C}_w(\bar{t} - \bar{\tau}(\bar{x}, \bar{y}))$. New variables appearing here include the dimensionless interstitial solute concentration, $\bar{C}_s = C_s(\bar{x}, \bar{y}, \bar{t})/C_0$, where $C_s(\bar{x}, \bar{y}, \bar{t})$ [$M L^{-3}$] is the solute mass per unit volume of interstitial fluid (i.e., as opposed to per bulk sediment volume) and $\bar{\tau}(\bar{x}, \bar{y}) = \tau(\bar{x}, \bar{y})/t_T$ is a dimensionless form of the location-dependent residence time function, $\tau(\bar{x}, \bar{y})$ [T], defined as the time it takes interstitial water parcels to travel from the SWI to any

(\bar{x}, \bar{y}) position in the sediment bed (Azizian et al., 2015) (derivation in Supplemental Information, **Text S3**):

$$\bar{\tau}(\bar{x}, \bar{y}) = \frac{\cos^{-1}[\cos \bar{x} e^{-\bar{y}}] - \bar{x}}{2 \cos \bar{x} e^{-\bar{y}}}, \quad \bar{y} > 0, \quad -\pi/2 < \bar{x} < \pi/2 \quad (8a)$$

The unsteady solution for the interstitial concentration of a conservative solute in the streambed directly follows from this last result, where the time-dependent solute concentration in the water column, $\bar{c}_w(\bar{t})$, is given by equation (5a):

$$\bar{c}_s(\bar{x}, \bar{y}, \bar{t}) = \begin{cases} \bar{c}_w(\bar{t} - \bar{\tau}(\bar{x}, \bar{y})), & \bar{t} \geq \bar{\tau}(\bar{x}, \bar{y}) \\ 0, & \bar{t} < \bar{\tau}(\bar{x}, \bar{y}) \end{cases}, \quad \bar{t} > 0, \quad \bar{y} > 0 \quad (8b)$$

It should be noted that equation (8a) is valid only within the bounds of the unit cell illustrated in **Figure 1c** (i.e., $-\pi/2 < \bar{x} < \pi/2$). Outside of the unit cell the equation must be translated, with or without reflection, using the following substitution rule for the dimensionless horizontal coordinate: $\bar{x} \rightarrow (-1)^n (\bar{x} - n\pi)$, where the integer n is given by $n = R[\bar{x}/\pi]$ and the function $R[\cdot]$ rounds to the nearest positive or negative integer value. Finally, a solution for the location of the concentration front in the sediment bed at any dimensionless time \bar{t} can be obtained by substituting \bar{t} for $\bar{\tau}$ on the left hand side of equation (8a), and numerically solving the resulting implicit expression for \bar{y} given \bar{x} , or vice versa. The implicit solution for the concentration front also applies to locations outside of the unit cell (i.e., for $\bar{x} < -\pi/2$ or $\bar{x} > \pi/2$) after translation with or without reflection, using the substitution rule presented above for \bar{x} .

3. Diffusive Model of Hyporheic Exchange by Bedform Pumping

Advection models, like the BPM, are premised on the idea that pore-scale advection dominates the transport of solutes in the hyporheic zone. Over the years, researchers have also explored the possibility of employing diffusive models to describe hyporheic exchange, generally, and bedform pumping, in particular (O'Connor and Harvey (2008)). EB, for example, argued that a diffusivity for bedform pumping should take the form of a dispersion coefficient, $E \approx 0.04 \lambda u_m / \theta$ [$L^2 T^{-1}$], where u_m / θ is a characteristic pore-scale velocity associated with the BPM (see equation (R5) in **Figure 1**) and the mixing length-scale is the bedform wavelength, λ (from equation (11) in Elliott and Brooks (1997b)). Applied to mass transfer in recirculating flumes, the constant diffusivity model predicts that the water column solute concentration, and the penetration depth of solute into the streambed, should both scale with the square root of time (Elliott and Brooks (1997a,b)). In their recirculating flume experiments, EB found that mass transfer across the SWI followed the predicted square root dependence until a transition time of around, $t_c \approx 8 \lambda \theta / u_m$. Afterwards, measured mass transfer rates fell below those predicted by the constant diffusivity model. Similarly, Marion and Zaramella (2005) reported that constant diffusivities inferred from recirculating flume studies decline as the timescale over which hyporheic exchange is measured increases.

From a mechanistic perspective, all of these problems with the constant diffusivity model can be rationalized by noting that, as time increases, mass transfer across the SWI slows dramatically as the relative contribution of deeper streamlines to bedform pumping increases (i.e., streamlines with starting positions in the range, $\bar{x}_{0c} < \bar{x}_0 < \pi/2$, see **Section 1** of SI). We hypothesize that this effect can be represented by requiring the dispersion coefficient to decline exponentially with depth, in keeping with the exponentially declining velocity field that underpins hyporheic exchange by bedform pumping (see equations (R2) - (R4) in **Figure 1**).

3.1. Governing Equations for Diffusive Bedform Pumping in a Closed System

Bedform pumping generates concentration fields in the interstitial fluids of the sediment bed that are at least two-dimensional (e.g., for artificially shaped triangular bedforms in the laboratory) and more often three-dimensional (e.g., in natural streams). However, if the goal is to predict average rates of mass transfer over, for example, a stream reach or a recirculating flume, knowledge of the two- and three-dimensional flow and subsurface concentration fields are not required. Thus, for many applications, mass transport and mixing by bedform pumping in the benthic biolayer can be conceptualized as an unsteady one-dimensional diffusion problem, for which the horizontally averaged vertical flux, $J(y,t)$ [$\text{M L}^{-2} \text{T}^{-1}$], of solute through the sediment is described by a flux-gradient diffusive model where the mixing coefficient, or effective diffusivity $D_{\text{eff}}(y)$ [$\text{L}^2 \text{T}^{-1}$], varies with depth through the benthic biolayer:

$$J(y,t) = -D_{\text{eff}}(y) \frac{\partial(\theta C_s)}{\partial y} \quad (9a)$$

Grant et al. (in review) demonstrated that equation (9a) is a reasonable descriptor of vertical solute transport by turbulent pumping through the benthic biolayer of a flat streambed, provided that the diffusion coefficient declines exponentially through the sediment bed. In this paper we hypothesize that a similar result applies to bedform pumping, but with the effective diffusivity replaced by an exponentially declining dispersion coefficient, $D_{\text{eff}}(y) = E(y) = E_0 e^{-ay}$. The surficial dispersion coefficient at the SWI, E_0 [$\text{L}^2 \text{T}^{-1}$], and the inverse decay length-scale, a [L^{-1}], are emergent properties of the two and three dimensional flow and concentration fields that drive bedform pumping; i.e., they are determined by spatial correlations between the time-averaged vertical component of the velocity field and the local mean solute concentration (Voermans et al., 2017). The corresponding one-dimensional mass balance equation can be written as follows:

$$\frac{\partial}{\partial t}(\theta C_s) = \frac{\partial}{\partial y} \left(E(y) \frac{\partial(\theta C_s)}{\partial y} \right) \quad (9b)$$

Equation (9b) equates the accumulation of mass in a differential horizontal slice of the sediment bed (left hand side) to the vertical diffusive transport (right hand side) of a conservative (non-reactive and non-adsorbing) solute (Incropera et al., 2007). The coordinate y increases with depth into the streambed and its origin (at $y=0$) is positioned at the horizontal plane of the SWI (**Figure 1b**).

Substituting the proposed functional form for the dispersion coefficient into equation (9b) and assuming streambed porosity θ does not vary substantially over the vertical dimension of the benthic biolayer (*ca.*, 5 cm) (Knapp et al., 2017), we arrive at the following mass balance equation for interstitial solute transport in the sediment bed:

$$\frac{\partial \bar{C}_s}{\partial \bar{t}} = e^{-\bar{y}} \frac{\partial^2 \bar{C}_s}{\partial \bar{y}^2} - e^{-\bar{y}} \frac{\partial \bar{C}_s}{\partial \bar{y}}, \quad \bar{y} > 0, \quad \bar{t} > 0 \quad (9c)$$

In equation (9c), time, $\bar{t} = t/t_E$, has been scaled by a characteristic timescale for dispersive mass transport through the benthic biolayer, $t_E = 1/(a^2 E_0)$, depth has been scaled by the inverse mixing length-scale, $\bar{y} = ay$, and the interstitial solute concentration has been scaled by the initial concentration in the overlying water column, $\bar{C}_s = C_s/C_0$ (same as for the BPM, see **Section 2.3**). By analogy to the BPM, we also assume that the streambed is initially solute free (equation (9d)), solute concentration drops off to zero deep in the streambed (equation (9e)), and the interstitial solute concentration at the top of the streambed equals the solute concentration in the overlying water column (equation (9f)) where $H(\bar{t})$ [-] is the Heaviside function (included here to satisfy the requirements of Duhamel's Theorem described later):

$$\bar{C}_s(\bar{y}, \bar{t}=0)=0 \quad (9d)$$

$$\bar{C}_s(\bar{y} \rightarrow \infty, \bar{t})=0 \quad (9e)$$

$$\bar{C}_s(\bar{y}=0, \bar{t})=\bar{C}_w(\bar{t})H(\bar{t}), \quad H(\bar{t})=\begin{cases} 0, & \bar{t} < 0 \\ 1, & \bar{t} > 0 \end{cases} \quad (9f)$$

In writing equation (9f) we have assumed that the interstitial concentration at the SWI is equal to the solute concentration in the overlying water column, which implies that mass transfer into the streambed is not rate-limited by convective mass transfer across the concentration boundary layer above the streambed; i.e., the dimensionless Biot Number—the ratio of timescales for diffusive mixing in the streambed and convective mass transfer across the turbulent boundary layer above the streambed—is much greater than unity (Incropera et al., 2007).

For a closed system with a well-mixed water column, like the recirculating flume illustrated in **Figure 3a**, mass balance over the water column takes the following form:

$$A_b h_w \frac{dC_w}{dt} = A_b \theta E_0 \left. \frac{\partial C_s}{\partial y} \right|_{y=0,t} \quad (10a)$$

In this equation, the change of solute mass in the overlying water column and recirculation system of the flume (left hand side) equals the mass transfer rate across the SWI by bedform pumping (represented here as a dispersive process, right hand side). Streambed porosity θ appears on the right-hand side of the equation to account for the abrupt change in area over which solute mass transport occurs above and below the SWI (Grant et al., 2012). Expressing equation (10a) using the dimensionless variables introduced above for the diffusion equation, we obtain equation (10b) where \bar{h}_w is a scaled form of the effective water depth.

$$\frac{d\bar{C}_w}{d\bar{t}} = \frac{1}{\bar{h}_w} \left. \frac{\partial \bar{C}_s}{\partial \bar{y}} \right|_{\bar{y}=0, \bar{t}} \quad (10b)$$

$$\bar{h}_w = \frac{a h_w}{\theta} \quad (10c)$$

3.2. Duhamel's Theorem and Green's Functions

As detailed in Grant et al. (in review), we can link the mass balance equations above (equation (10b)) and below (equation (9c)) the SWI, and thereby account for two-way coupling across the SWI, using Duhamel's Theorem, an analytical approach for solving the diffusion equation in cases where the forcing function at one boundary is a piece-wise continuous function of time (Perez Guerrero et al., 2013). Duhamel's Theorem allows us to express the evolution of interstitial solute concentration in the sediment bed as a convolution of the water column concentration $C_w(\bar{t})$ and a so-called auxiliary function $C_s^A(\bar{y}, \bar{t})$ where v is a dummy integration variable (Perez-Guerrero et al., 2013):

$$\bar{C}_s(\bar{y}, \bar{t}) = \int_0^{\bar{t}} \bar{C}_s^A(\bar{y}, \bar{t} - v) \frac{d}{dv} [\bar{C}_w(v) U(v)] dv \quad (11a)$$

The auxiliary function is a solution to the same system of equations (equations (9c), (9d), (9e), and (9f)), but with the inhomogeneous term replaced by a unit step function (compare equations (9f) and (11e)):

$$\frac{\partial \bar{C}_s^A}{\partial \bar{t}} = e^{-\bar{y}} \frac{\partial^2 \bar{C}_s^A}{\partial \bar{y}^2} - e^{-\bar{y}} \frac{\partial \bar{C}_s^A}{\partial \bar{y}} \quad (11b)$$

$$C_s^A(\bar{y}, \bar{t} = 0) = 0, \bar{y} \geq 0 \quad (11c)$$

$$C_s^A(\bar{y} \rightarrow \infty, \bar{t}) = 0 \quad (11d)$$

$$C_s^A(\bar{y} = 0, \bar{t}) = H(\bar{t}) \quad (11e)$$

Substituting the coordinate transformation, $\xi = e^{\bar{y}}$, into equation (11b) (Yates, 1992) and solving the resulting system of equations in the Laplace Domain, we arrive at the following analytical

385 solution for the auxiliary function where $\bar{s} = s t_e$ is a dimensionless form of the Laplace

386 Transform variable:

$$387 \quad \bar{C}_s^u(\bar{y}, \bar{s}) = \frac{\sqrt{e^{\bar{y}}} K_1(2\sqrt{\bar{s}e^{\bar{y}}})}{\bar{s} K_1(2\sqrt{\bar{s}})} \quad (12)$$

388 Duhamel's Theorem (equation (11a)) can also be expressed as a convolution of the

389 dimensionless water column concentration, $\bar{C}_w(\bar{t})$, and a so-called Green's function, $\bar{G}(\bar{y}, \bar{t})$ [T⁻¹],

390 which is scaled here by the dispersive mixing timescale introduced earlier, $\bar{G}(\bar{y}, \bar{t}) = t_e G(\bar{y}, \bar{t})$:

$$391 \quad \bar{C}_s(\bar{y}, \bar{t}) = \int_0^{\bar{t}} \bar{G}(\bar{y}, \bar{\tau}) \bar{C}_w(\bar{t} - \bar{\tau}) d\bar{\tau} \quad (13a)$$

$$392 \quad \bar{G}(\bar{y}, \bar{t}) = \frac{\partial \bar{C}_s^u}{\partial \bar{t}} \quad (13b)$$

393 Substituting equation (12) into equation (13b) yields a Green's function for the exponentially

394 declining diffusivity profile:

$$395 \quad \bar{G}_E(\bar{y}, \bar{t}) = \sqrt{e^{\bar{y}}} \mathcal{L}^{-1} \left[\frac{K_1(2\sqrt{\bar{s}e^{\bar{y}}})}{K_1(2\sqrt{\bar{s}})} \right] \quad (13c)$$

396 The similarity between equation (13a) and the convolution integral derived earlier for the BPM

397 (equation (1a)) is striking. In both cases, the water column concentration is convolved with a

398 function (Green's function in the case of the diffusive model and an RTD function in the case of

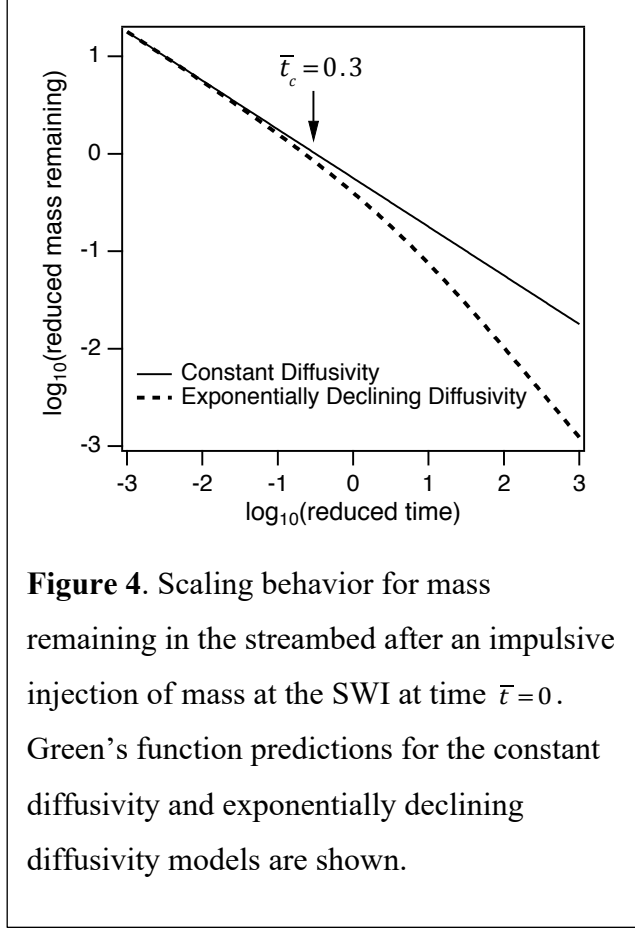
399 the advective model) that captures the response of the streambed to an impulsive injection of

400 mass into the SWI at dimensionless time $\bar{t} = 0$, $\bar{C}_w(\bar{t}) = \delta(\bar{t})$, where $\delta(\bar{t})$ [-] is the Dirac Delta.

401 The Green's function above (equation (13c)) is specific for the choice of an exponentially

402 decaying diffusivity profile. For the same set of initial and boundary conditions, a second

403 Green's function can be derived for a constant diffusivity profile (the so-called "Signaling



Problem”, see Gorenflo and Mainardi, 1988): $\bar{G}_c(\bar{y}, \bar{t}) = \bar{y} e^{-\bar{y}^2/(4\bar{t})} / (2\bar{t} \sqrt{\pi \bar{t}})$. With these two Green's functions we can interrogate how the choice of a diffusivity profile (i.e., exponentially declining or constant) influences the temporal scaling of mass remaining in the sediment bed following an impulsive input of mass at the SWI at time $\bar{t} = 0$. This is because, for $\bar{t} > 0$, the upper boundary condition for these two Green's functions is zero, $\bar{C}_w(\bar{t} > 0) = \delta(\bar{t} > 0) = 0$, and therefore solute

416 mass in the sediment bed, $\bar{m} = \int_0^\infty \bar{G}(\bar{y}, \bar{t}) d\bar{y}$,

417 will diffuse back into the water column after its injection at time $\bar{t} = 0$:

418
$$\bar{m}_E(\bar{t}) = \mathcal{L}^{-1} \left[\frac{K_0(2\sqrt{s})}{\sqrt{s} K_1(2\sqrt{s})} \right] \quad (14a)$$

419
$$\bar{m}_c(\bar{t}) = \frac{1}{\sqrt{\pi \bar{t}}} \quad (14b)$$

420 As might be expected based on the discussion at the beginning of **Section 3**, the constant
 421 diffusivity model predicts solute mass remaining in the sediment bed declines inversely with the
 422 square root of time (equation (14b)). The exponentially declining diffusivity model (equation
 423 (14a)) exhibits this same $1/\sqrt{\bar{t}}$ scaling initially, but falls off more rapidly after $\bar{t}_c = t_c/t_E \approx 0.3$

(**Figure 4**). This result can be rationalized by noting that, for an exponentially declining diffusivity profile, deeper portions of the streambed are relatively inaccessible to solute injected at the SWI at $\bar{t}=0$ and consequently contribute little to the release of stored mass at long times. The similarity between the scaling behavior illustrated in **Figure 4** and the scaling behavior described earlier for mass transfer across the SWI in recirculating flumes (see preamble to **Section 3**) is the first indication that our overarching hypothesis—that bedform pumping can be represented by an exponentially decaying diffusivity model—may be valid.

3.3. Solute Concentration in the Water and Sediment Columns of Closed System

From the results presented above, a set of explicit solutions can be derived for solute concentration in the water column and interstitial fluids of a closed system (Grant et al., in review):

$$\bar{c}_w(\bar{t}) = \mathcal{L}^{-1} \left[\frac{1/\bar{s}}{1 - \frac{1}{\bar{s}h_w} \left(\partial \tilde{G} / \partial \bar{y} \right)_{\bar{y}=0, \bar{s}}} \right] \quad (15a)$$

$$\bar{c}_s(\bar{y}, \bar{t}) = \mathcal{L}^{-1} \left[\frac{\tilde{G}(\bar{y}, \bar{s})/\bar{s}}{1 - \frac{1}{\bar{s}h_w} \left(\partial \tilde{G} / \partial \bar{y} \right)_{\bar{y}=0, \bar{s}}} \right] \quad (15b)$$

These analytical solutions are written in terms of the Laplace transform of the Green's function and its derivative, which, in this context, are tailored to the choice of diffusivity profile. For an exponentially declining diffusivity profile, they are as follows:

$$\tilde{G}_E(\bar{y}, \bar{s}) = \sqrt{e^{\bar{y}}} \frac{K_1(2\sqrt{\bar{s}}e^{\bar{y}})}{K_1(2\sqrt{\bar{s}})} \quad (15c)$$

$$\left. \frac{\partial \tilde{G}_E}{\partial \bar{y}} \right|_{\bar{y}=0, \bar{s}} = - \frac{\sqrt{\bar{s}} K_0(2\sqrt{\bar{s}})}{K_1(2\sqrt{\bar{s}})} \quad (15d)$$

For a constant diffusivity profile, these two functions can be computed directly from the solution to the Signaling Problem introduced earlier:

$$\tilde{G}_c(\bar{y}, \bar{s}) = e^{-\bar{y}\sqrt{\bar{s}}} \quad (15e)$$

$$\left. \frac{\partial \tilde{G}_c}{\partial \bar{y}} \right|_{\bar{y}=0, \bar{s}} = -\sqrt{\bar{s}} \quad (15f)$$

4. Experimental Evaluation of Advective and Diffusion Models of Bedform Pumping

4.1. EB's Bedform Pumping Dataset and Model Optimization

To test the parallel advective and diffusive analytical frameworks derived above, we turned to one of the first published recirculating flume experiments specifically designed, along the lines of **Figure 3a**, to investigate the unsteady transfer of a conservative solute across the SWI by bedform pumping (Elliott and Brooks, 1997b). EB's experiments were conducted with stationary bedforms (either artificial triangular bedforms or natural ripples), a non-adsorbing and stable fluorescent dye (Lissamine), and under various flow velocities (8.6 to 13.2 cm s⁻¹), water depths (1.14 to 2.54 cm) and shear velocities (1.3 to 3 cm s⁻¹) (Experiment ID's 8, 9, 12, 14 – 17). The sediment bed, which ranged in depth from 12.5 to 22.0 cm depending on the experiment, consisted of medium or fine-grained unconsolidated sand of hydraulic conductivity $K_h = 0.11$ and 0.0079 cm s⁻¹ and porosity $\theta = 0.325$ and 0.295, respectively (a summary of experimental conditions is included in the Supplemental Information, **Table S1**). Published over 20 years ago, EB's study remains one of the few where the evolution of dye concentration is followed both above and below the SWI—a feature we will take advantage of here.

Experimental evaluation of our analytical models was carried out in two steps. First, we fit the advective and diffusive models for $c_w(t)$ (equations (5a) and (15a), respectively) to EB's

measurements of dye concentration in the water column over time. This was accomplished using the NonLinearModelFit routine in the Mathematica computing package (v.12, Wolfram Research, Inc.) implemented on UC Irvine's High-Performance Computing Cluster. Laplace inversions were carried out by Gaussian Quadrature in the Mathematica package authored by U. Graf (Graf, 2004). This fitting exercise yielded, for each of EB's experiments, inferred values for the half-amplitude of the pressure head variation and effective water depth (h_m and h_w , advective model) and the surficial dispersion coefficient and inverse mixing length-scale (E_0 and a , diffusive model), together with the standard deviation of each parameter and the model's coefficient of variation (R^2). For consistency, h_w values inferred from the advective model were applied to the diffusive model; all other parameters (θ , K_h , λ , and C_0) were reported by EB for each experiment (see **Table S1**). In the second step, parameter values inferred from the water column studies were used to predict the movement of dye through the interstitial fluids of the streambed over time (equations (8b) and (15b)). These model predictions were compared to observations of the dye front in the sediment bed, which EB recorded by periodically marking the location of the leading edge of the dye plume on the side of their flume (the wall of the flume was transparent, and dye was visualized with a hand-held UV light).

4.2. Evaluation of Model-Predicted Water Column Solute Concentrations

Across all seven experiments, the advective model (equation (5a)) and diffusive model (with an exponentially declining diffusivity profile, equation (15a)) closely conform to EB's time series measurements of dye concentration in the water column ($R^2 > 0.9998$ for both models, **Figure 5a,b**, also see **Tables S2** and **S3** in the Supplemental Information). For comparison, water column concentrations for Experiment #17 were also simulated with the constant diffusivity model; this involved substituting equation (15f) into equation (15a) and adopting the superficial

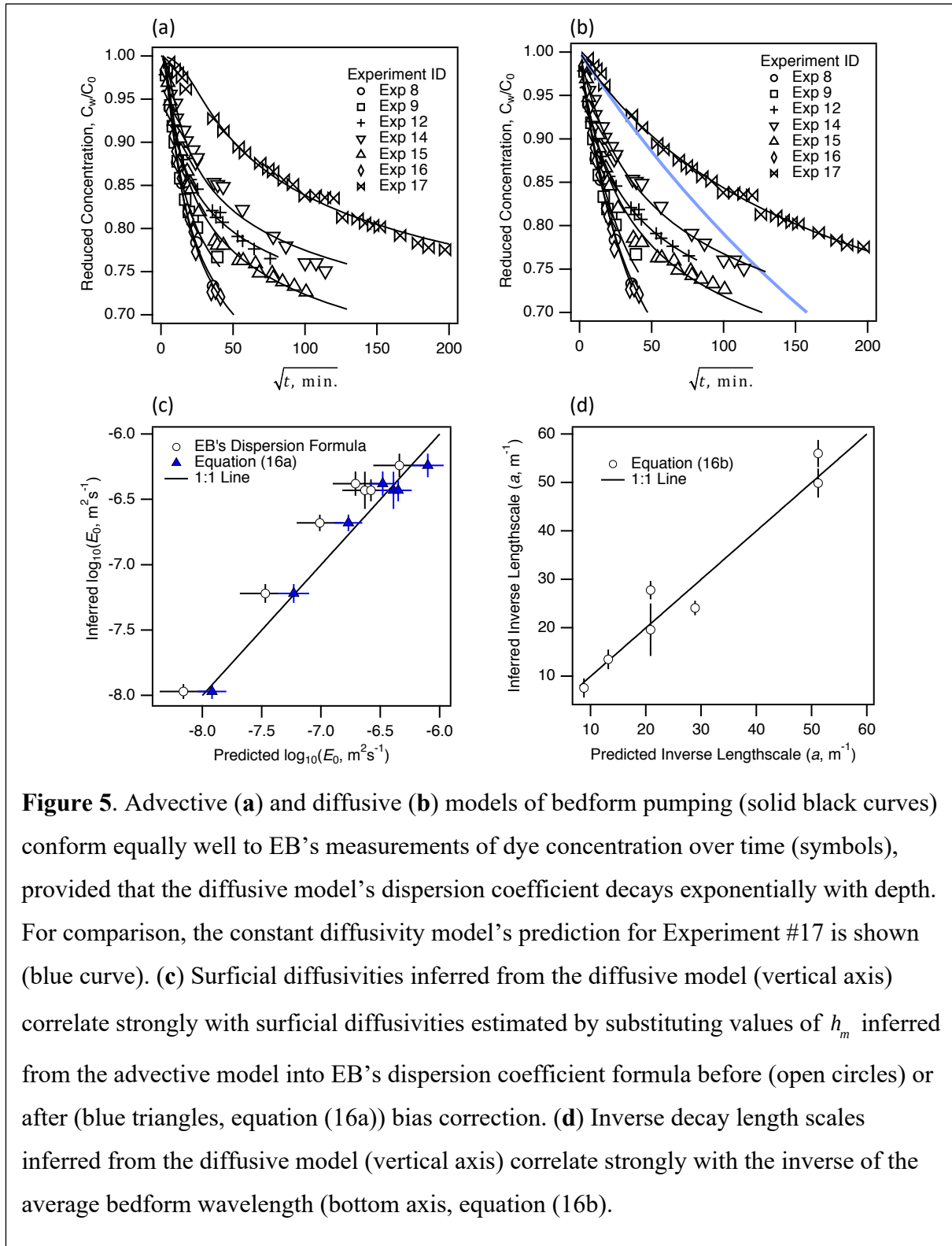


Figure 5. Advective (a) and diffusive (b) models of bedform pumping (solid black curves) conform equally well to EB's measurements of dye concentration over time (symbols), provided that the diffusive model's dispersion coefficient decays exponentially with depth. For comparison, the constant diffusivity model's prediction for Experiment #17 is shown (blue curve). (c) Surficial diffusivities inferred from the diffusive model (vertical axis) correlate strongly with surficial diffusivities estimated by substituting values of h_m inferred from the advective model into EB's dispersion coefficient formula before (open circles) or after (blue triangles, equation (16a)) bias correction. (d) Inverse decay length scales inferred from the diffusive model (vertical axis) correlate strongly with the inverse of the average bedform wavelength (bottom axis, equation (16b)).

diffusivity, E_0 , inferred from fitting the exponentially declining diffusivity model to the same dataset (Table S3). The constant diffusivity model also conforms to measurements of the water

column concentration until around $\sqrt{t} \approx \sqrt{30} \text{ min}^{1/2}$; thereafter, the constant diffusivity model seriously underpredicts observed concentration measurements (blue curve, **Figure 5b**). Water column concentrations predicted by the constant diffusivity model decline approximately linearly when plotted against \sqrt{t} , consistent with EB's observations about the constant diffusivity model (see preamble to **Section 3**) and the \sqrt{t} -scaling of the constant diffusivity's Green's function (see equation (14b) and **Figure 4**).

We can also evaluate the advective and diffusive models based on how well their inferred parameter values reproduce values expected based on theory or measurements. For example, values of the half-amplitude pressure head inferred from the advective model (ranging from $h_m = 0.04$ to 0.57 mm) are similar (roughly factor of two or better) to values estimated from EB's empirical formula (equation (7)) (ranging from $h_m = 0.09$ to 0.31 mm , **Table S2**). Likewise, values of the effective water depth inferred from the advective model (ranging from $h_w = 8.8$ to 16.7 cm) are similar (roughly factor of two or better) to values estimated from reported flume water volume (excluding interstitial fluid) and streambed area ($h_w = V_w / A_b = 11.3$ to 12.5 cm) (**Table S2**). Deviations between inferred and predicted (or measured) values of h_m and h_w do not necessarily imply that the model-generated values are incorrect. For example, the half-amplitude head values predicted by equation (7) are only approximately correct (Shen et al., 1990; Fox et al., 2014). Measurement errors associated with flume water volume and bed surface area (which may be difficult to define, given the undulatory nature of the SWI with bedforms) also contribute uncertainty and bias to experimental estimates of h_w .

A more rigorous assessment of the inferred parameter values can be framed as follows:

Can parameter values inferred from the advective model be translated directly into parameter

values for the diffusion model and vice versa? To answer this question, we equated EB's proposed formula for a bedform pumping dispersion coefficient to the diffusivity model's surficial dispersion coefficient: $E_0 \approx 0.04\lambda u_m / \theta$. Substituting the BPM's solution for the maximum Darcy flux (u_m , equation (R5) in **Figure 1**), this formula predicts that the diffusive model's surficial dispersion coefficient is directly proportional to the advective model's half-amplitude pressure head, $E_0 \approx 0.08\pi K_h h_m / \theta$. When values of h_m inferred from the advective model are substituted into this formula, the predicted values of E_0 are highly correlated with values of E_0 inferred from the diffusion model (Pearson's Correlation coefficient, $R=0.99$) (open black circles, **Figure 5c**). Adjusting the equation's pre-factor to correct the bias evident in the figure, we arrive at the following relationship between the advective and dispersive descriptions of bedform pumping (blue filled triangles, **Figure 5c**):

$$E_0 \approx 0.133\pi K_h h_m / \theta, \quad 0.08 \leq K_h [mm s^{-1}] \leq 1.1, \quad 0.042 \leq h_m [mm] \leq 0.11, \quad 0.295 \leq \theta \leq 0.325 \quad (16a)$$

Likewise, the inverse decay length-scale, a , inferred from the diffusion model is highly correlated ($R^2=0.95$) with the inverse of the average bedform wavelength (**Figure 5d**):

$$a [cm^{-1}] = 5.28 / \lambda [cm] - 0.0882, \quad 8.8 \leq \lambda [cm] \leq 30 \quad (16b)$$

Equations (16a) and (16b) provide a direct link between our advective and diffusive descriptions of bedform pumping, such that a parameter set for one can be directly translated into a parameter set for the other. The implication is that these two descriptions of bedform pumping are, in fact, functionally equivalent, provided that the limitations with existing analytical models outlined earlier (water column concentration constant in time, diffusivity is constant in depth, two-way coupling across the SWI neglected, and the sediment bed's finite porosity neglected) are properly addressed, as they have been in this study. Because equations (16a) and (16b) are calibrated with

data from EB's study alone, they are per force restricted to a limited range of flow and streambed conditions (indicated by the inequalities above). A meta-analysis is underway to evaluate the predictive power of these equations beyond the set of experiments analyzed here.

4.3. Evaluation of Model-Predicted Interstitial Solute Concentrations

We can also evaluate the advective and diffusive models relative to their ability to predict the unsteady transport of dye plumes through the interstitial fluids of the sediment bed. The progression of one such plume beneath an artificial triangular ripple (EB's Experiment #9) is reproduced in **Figures 6a - d** (thick dashed curves). The dye plume penetrated to a depth of about 8 cm in the first 75 minutes, but required an additional 575 minutes to progress downward another 4 cm. To compare these observations with the advective model solution, the BPM's coordinate system must first be aligned with EB's triangular ripple. To this end we used the parameter values estimated from the water column optimization study of Experiment #9 (see **Table S2**) to predict (with equation (8b)) the interstitial dye concentration in the sediment bed at $t = 75$ minutes, coinciding with EB's first dye front measurement. The model's horizontal coordinate was adjusted to align the left and right edges of the observed and predicted dye fronts. Finally, the model's vertical coordinate was adjusted so that the top of the (flat) model domain is equidistant between the crest and trough of the triangular bedform (final alignment is shown in **Figure 6a**). After making these adjustments, the advective model's predictions for the downward migration of the dye plume over time closely agree with EB's observations of the dye front at $t = 150, 320, \text{ and } 650$ minutes (**Figures 6b - 6d**).

Two-way coupling is also evident in the model-predicted interstitial concentration field. Predicted dye concentrations are elevated along the front of the plume because water parcels at the front moved into the sediment bed near time $t = 0$ when dye concentration in the water

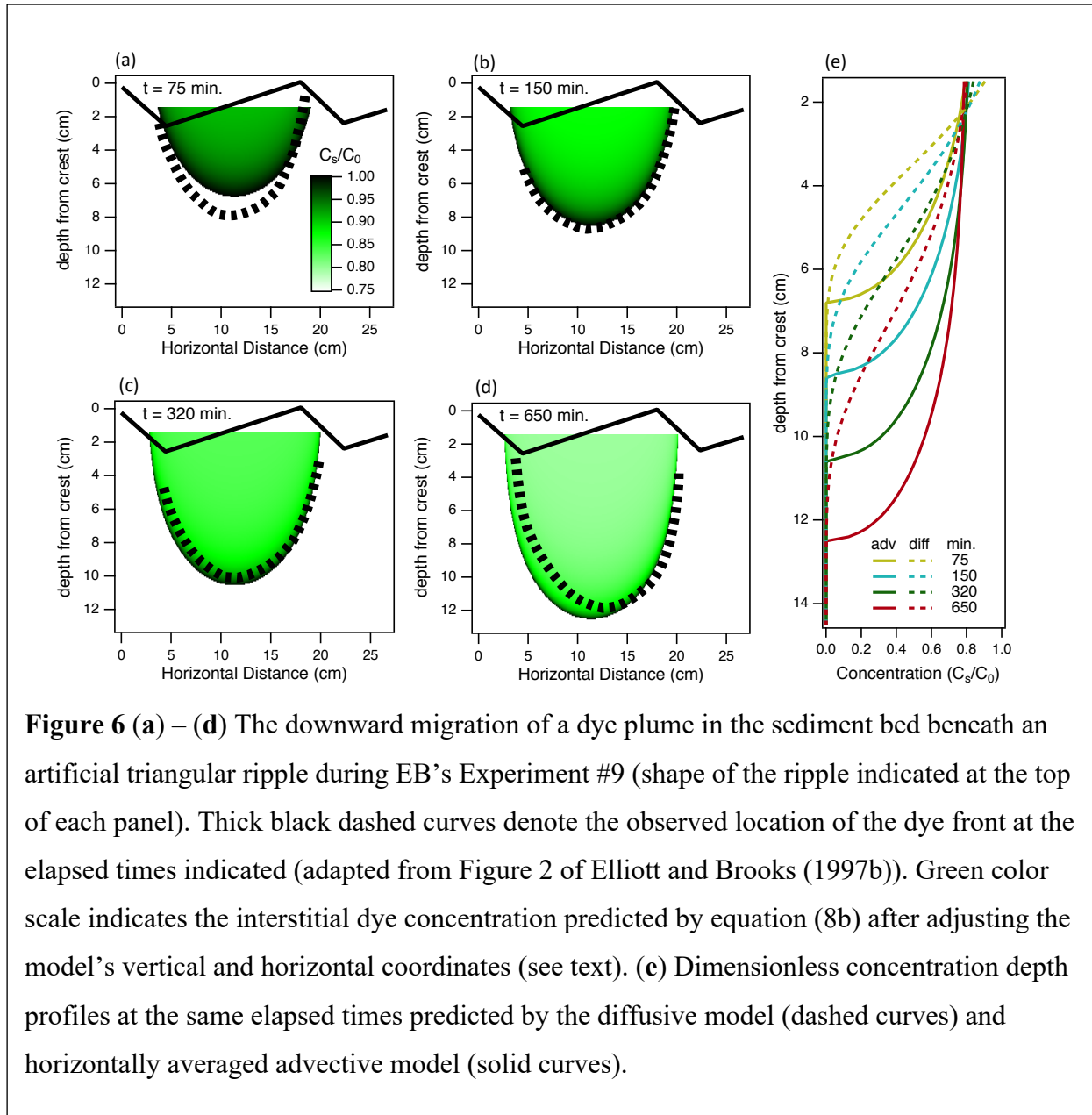


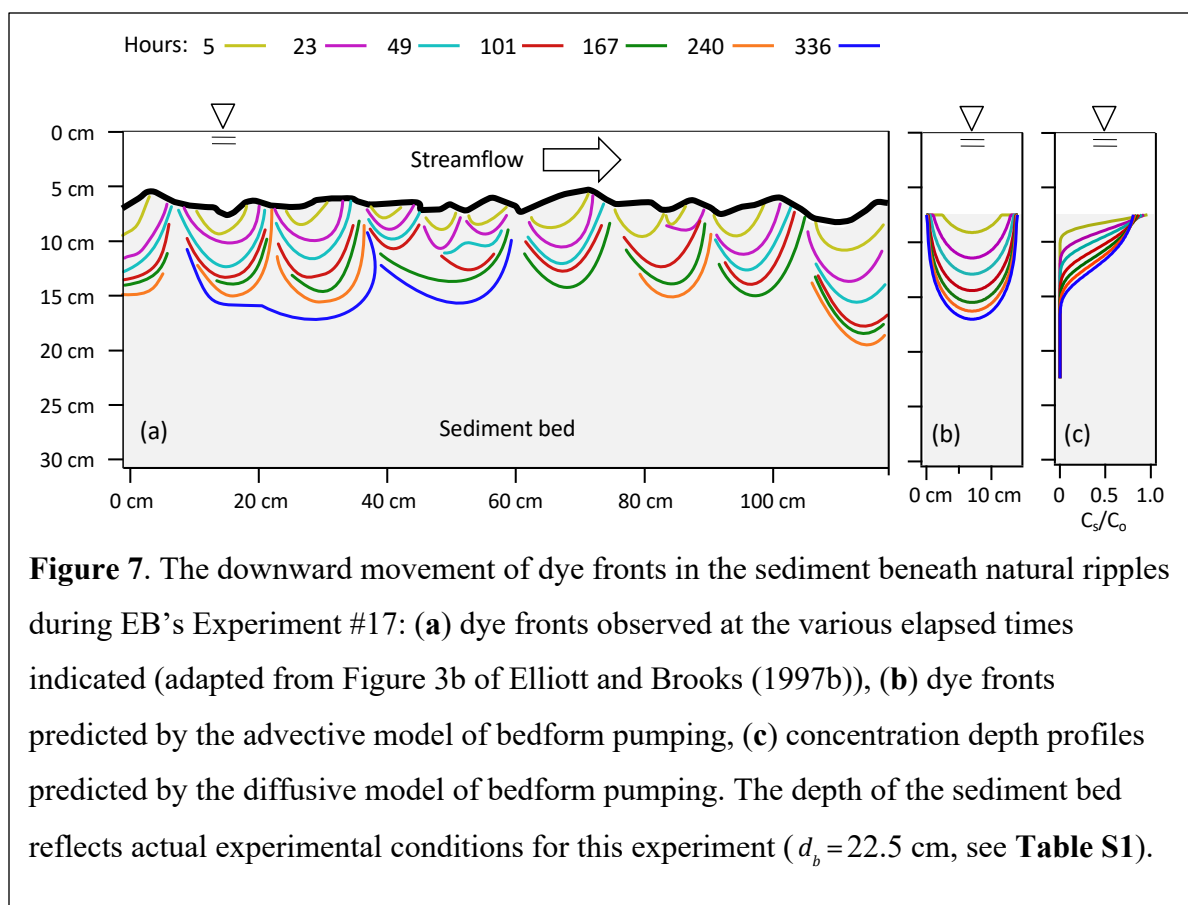
Figure 6 (a) – (d) The downward migration of a dye plume in the sediment bed beneath an artificial triangular ripple during EB’s Experiment #9 (shape of the ripple indicated at the top of each panel). Thick black dashed curves denote the observed location of the dye front at the elapsed times indicated (adapted from Figure 2 of Elliott and Brooks (1997b)). Green color scale indicates the interstitial dye concentration predicted by equation (8b) after adjusting the model’s vertical and horizontal coordinates (see text). (e) Dimensionless concentration depth profiles at the same elapsed times predicted by the diffusive model (dashed curves) and horizontally averaged advective model (solid curves).

column was near its initial (maximum) value, $C_w = C_0$. As time progresses, dye concentration in the water column declines and the trailing edge of the dye plume, which consists of younger water parcels, becomes less concentrated. This pattern—high dye concentration along the plume’s front and low concentration along the plume’s trailing edge—is particularly striking for the simulation at $t = 150$ (**Figure 6b**). Eventually the plume’s concentration field takes on a

more uniform appearance as older water parcels (with higher dye concentrations) return to the stream along slow moving streamlines (**Figure 6d**).

Thus far we have found little difference between our advective and diffusive models of bedform pumping. One aspect where these two models differ substantially is their respective concentration depth profiles (**Figure 6e**). The diffusivity model's depth profiles are convex in shape and characterized by a diffuse concentration front that becomes increasingly smeared out over the vertical extent of the streambed with increasing time. By contrast, the advective concentration depth profiles (generated by horizontally averaging the two-dimensional concentration fields appearing in **Figures 6a – 6d**) are convex and characterized by persistent and very sharp concentration fronts. These contrasting shapes, which reflect the purely advective and diffusive transport mechanisms underlying the two modeling frameworks, could lead to very different predictions for the transport of reactive solutes through the benthic biolayer. This begs the question: *which of these two profiles is more representative of natural systems?*

To answer this question, we turned to recirculating flume experiments EB conducted with stationary natural ripples. These experiments entailed operating the flume under high flow conditions (to induce sediment transport and ripple formation) and then lowering the flow velocity (to immobilize the bedforms and conduct the dye exchange experiments). Not surprisingly, dye plumes generated by natural ripples are variable with respect to their horizontal extent and the depth to which dye penetrates the streambed (**Figure 7a**). This variability, which arises from variations in bedform geometry (i.e., height H and wavelength λ) and the three-dimensional nature of natural ripples, can be formally analyzed using spectral methods (e.g., Stonedahl et al. (2010)). However, if the goal is to obtain bulk estimates for the downward



progression of solute through the benthic biolayer over time, both the advective and diffusive analytical solutions derived in this study perform remarkably well (compare **Figures 7a – 7c**). The sharp dye fronts predicted by the advective model are comparable to patterns of dye penetration beneath “average” bedforms; e.g., the two dye plumes located 10 to 40 cm along the horizontal axis (**Figure 7a**). The smeared-out dye fronts predicted by the diffusive model, on the other hand, may be more representative of the concentration profile one would obtain by horizontally averaging the interstitial concentration field across all bedforms (this hypothesis could not be tested with EB’s dataset because these authors recorded the time evolution of concentration fronts, not concentration fields). What these results imply for reactive solute transport through the benthic biolayer is an interesting topic for future study.

5. Discussion

The functional equivalence of the analytical advective and diffusive frameworks derived here implies that their application can be tailored to the problem at hand. The advective model is a relatively faithful representation of the two-dimensional interstitial flow fields associated with bedform pumping. Consequently, this framework will be useful in cases where knowledge of flow paths through the benthic biolayer, and their associated Darcy fluxes and residence times, is required. The removal of stream borne particles in the benthic biolayer by deep bed filtration, for example, requires detailed information about the interstitial flow field. This is because, as particles move through the streambed, their filtration rate depends on the local flow velocity (through the contact efficiency η_0 [-], see Tufenkji and Elimelech (2004)) which varies continuously along a streamline (see equations (R1)-(R5) in **Figure 1**). Another example is the spatial zonation of interstitial oxygen concentration beneath bedforms, including the formation of so-called “anoxic chimneys” in upwelling zones (Kessler et al., 2012; Kessler et al., 2013; Azizian et al., 2015). This biogeochemical zonation, which arises from the coupling between in-bed redox reactions and bedform pumping of electron donors and acceptors, can impose significant constraints on important streambed functions, such as coupled nitrification-denitrification (Kessler et al., 2013). Because the advective model’s flow field is quantitatively linked to bedform geometry and stream flow (i.e., bedform height and wavelength, as well as stream depth and velocity, see equation (7)), physicochemical (e.g., particle filtration) and biogeochemical (e.g., nutrient transformation) functions of the benthic biolayer can be tied directly to geomorphic processes, such as the adjustment of bedform morphology to changes in land use and flow regime, for example as a result of urbanization (Harvey et al., 2012).

On the other hand, a strength of the diffusive model is its ability to combine multiple mechanisms for mass transport across the SWI. As noted earlier, mixing across a flat SWI can be

616 characterized by an effective diffusivity, D_{eff} , that incorporates three transport mechanisms
 617 (Richardson and Parr, 1988; O'Connor and Harvey, 2008; Grant et al., 2012; Voermans et al.,
 618 2018): (1) tortuosity-modified molecular diffusion (D_m [$\text{L}^2 \text{T}^{-1}$]); (2) dispersion (D_d [$\text{L}^2 \text{T}^{-1}$]); and
 619 (3) turbulent diffusion (D_t [$\text{L}^2 \text{T}^{-1}$]). The turbulent and dispersive diffusivities increase with the
 620 Permeability Reynolds Number, $\text{Re}_K = u_* \sqrt{K} / \nu$ [-], a dimensionless ratio of a permeability length
 621 scale (\sqrt{K} [L]) and the viscous length scale that governs turbulence at the surface of the
 622 streambed (ratio of the kinematic viscosity of water ν [$\text{L}^2 \text{T}^{-1}$] and the shear velocity u_* [L T^{-1}])
 623 (Voermans et al., 2017; Voermans et al., 2018). For turbulent mass transfer across a flat SWI, and
 624 accounting for the exponential decay of diffusivity with depth, the surficial effective diffusivity
 625 exhibits different Permeability Reynolds Number scaling behavior in the dispersive ($D_{\text{eff},0} \propto \text{Re}_K^{2.5}$,
 626 $0.01 < \text{Re}_K < 1$) and turbulent diffusive ($D_{\text{eff},0} \propto \text{Re}_K^1$, $\text{Re}_K > 1$) regimes (Grant et al., in review). Our
 627 formula linking advective and dispersive descriptions of bedform pumping (equation (16a))
 628 implies that dispersive mixing by bedform pumping also increases with the Permeability
 629 Reynold Number, $E_0 \propto \text{Re}_K^2$. This last result can be demonstrated by substituting into equation
 630 (16a) definitions for the Darcy-Weisbach friction factor, $f_D = 8u_*^2 / V^2$ [-] (Sabersky and Acosta,
 631 1989) and the streambed permeability, $K = \nu K_h / g$ [L^2] (McCabe et al., 2010), and noting that the
 632 friction factor will likely scale with the fraction of the water depth taken up by a bedform,
 633 $f_D \propto (H/d)^\gamma$. Thus, the one-dimensional diffusive framework potentially solute mixing through the
 634 benthic biolayer by molecular diffusion, turbulent dispersion, turbulent diffusion, and bedform
 635 pumping.

Another advantage of the diffusive model is that it can be readily modified to account for groundwater recharge or discharge (through the addition of an advective term to equation (9a)) and the inclusion of biogeochemical reaction networks, such as the Monod kinetic expressions associated with nitrogen cycling in streambeds (respiration, ammonification, nitrification, and denitrification (Azizian et al., 2017)). While surface water-groundwater exchange can be factored into the BPM's flow field as well (c.f., Boano et al., 2008) doing so invalidates a key requirement of the advective model's predictions for solute transport; namely, that the x -component of the velocity is everywhere constant along a streamline (see **Text S1** in Supplemental Information). While outside of the scope of this paper, it is also interesting to note that, at sufficiently high celerity, bedform migration tends to reduce the complexity of the interstitial concentration fields, in effect transforming the complex two- and three-dimensional concentration fields associated with bedform pumping across stationary bedforms into simple one-dimensional vertical concentration gradients (e.g., of interstitial oxygen concentrations, Wolke et al., 2020) that may also be amenable to analysis with a diffusive modeling framework.

A benefit of analytical models (compared to numerical simulations) is the relative ease with which they can be implemented, and the physical insights afforded by expressing the quantity of interest (e.g., interstitial solute concentration or mass flux across the SWI) as an explicit function of key system variables. An obvious limitation is that their derivation often entails simplifying assumptions that may not be valid in practice. Key assumptions associated with the advective and diffusive modeling frameworks derived here include: (1) the interstitial flow field underlying bedform pumping is steady-state, although solute concentrations in the water column and interstitial fluids of the streambed may vary with time while fully accounting for two-way coupling across the SWI; (2) the solutions are specific to closed systems (such as

recirculating flumes) while most hyporheic exchange problems of practical interest are open systems (such as streams and coastal sediments); (3) the transport properties of the benthic biolayer (such as porosity and hydraulic conductivity) are assumed homogeneous and invariant with time while a previous study found that these assumptions can be invalidated by bioclogging (Caruso et al., 2017); and (4) as already noted, our results assume the solute in question is conservative (i.e., non-reactive and does not absorb to the porous matrix) while most hyporheic exchange problems of practical interest involve reactive solutes or particles. These limitations can be addressed, to varying degrees, within the context of our analytical framework, and efforts to do so are currently underway.

6. Conclusions

In this paper we derived two parallel analytical frameworks, one advective and the other diffusive, that together relax many of the assumptions that limit the practical utility of presently available analytical models for bedform pumping. Both frameworks allow the water column concentration to vary with time while accounting for the two-way coupling of solute concentrations above and below the SWI; the diffusive framework additionally allows the mixing rate, or diffusivity, to vary with depth through the sediment bed. When applied to previously published measurements of bedform pumping in a recirculating flume (Elliott and Brooks, 1997a), we find that both analytical frameworks closely reproduce average patterns and rates of hyporheic exchange, *provided that the diffusion model's diffusivity declines exponentially with depth*. Practical application of these two frameworks can be tailored to the problem at hand, depending on whether detailed knowledge of the interstitial flow fields and associated Darcy fluxes and residence times is required (advective model) or solute transport across the SWI is subject to multiple transport mechanisms, not just bedform pumping (diffusive model). Because the advective framework is grounded in a physical description of bedform

pumping, it explicitly accounts for how changes in stream flow and sediment transport (e.g., associated with urbanization) influence bedform geometry (wavelength and height), the half-amplitude of the pressure head variation, and the hydraulic conductivity of the sediment bed. The exponentially declining diffusivity framework, on the hand, lumps these geomorphic processes into a surficial dispersion coefficient and an inverse decay length-scale that can be directly calculated from the aforementioned advective model parameters (see equation (16a,b)). The latter formula also predicts that the surficial dispersion coefficient for bedform pumping increases with the dimensionless Permeability Reynolds Number, consistent with diffusivities measured for turbulent exchange across flat streambeds (Voermans et al., 2018; Grant et al., in review) and streambeds with bedforms (O'Connor and Harvey, 2008; Grant et al., 2012; Grant et al., 2018). Efforts are currently underway to extend these analytical solutions to open systems (e.g., stream networks), bedform turnover, unsteady flows, and the non-linear reactions that drive nutrient cycling in the benthic biolayer of streams.

Acknowledgments and Data

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853 **Table 1.** A summary of the probability distributions trialed as potential descriptors of the
854 Bedform Pumping Model's residence time distribution and their inferred parameter values.

| Distribution Name | PDF for the BPM's dimensionless residence time, $f_{\text{RTD}}(\bar{t})$ | Inferred Parameter Values | Kolmogorov Smirnov Test | |
|-------------------|--|---------------------------------------|-------------------------|------|
| | | | Statistic | Rank |
| Fréchet | $\frac{\beta}{1-e^{-\beta/\mu}} \frac{e^{-\frac{\beta}{\mu+\bar{t}}}}{(\mu+\bar{t})^2}, \bar{t} \geq 0$ | $\beta=1.6, \mu=0.2$ | 0.00881 | 1 |
| Pareto | $\frac{\alpha k^{-1/\gamma}}{\gamma} (\bar{t}^{1/\gamma-1}) \left(1 + \left(\frac{k}{\bar{t}} \right)^{-1/\gamma} \right)^{-(1+\alpha)}, \bar{t} \geq 0$ | $k=1.137, \alpha=0.504, \gamma=0.557$ | 0.01088 | 2 |
| Log Normal | $\frac{e^{-\frac{(\ln \bar{t} - \mu)^2}{2\sigma^2}}}{\bar{t} \sqrt{2\pi\sigma^2}}, \bar{t} \geq 0$ | $\mu=0.891, \sigma=1.405$ | 0.05547 | 3 |
| Gamma | $\frac{\beta^{-\alpha}}{\Gamma[\alpha]} \bar{t}^{\alpha-1} e^{-\frac{\bar{t}}{\beta}}, \bar{t} \geq 0$ | $\alpha=0.267, \beta=126.7$ | 0.30746 | 4 |
| Exponential | $\lambda e^{-\lambda \bar{t}}, \bar{t} \geq 0$ | $\lambda=0.03$ | 0.62029 | 5 |

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