

1 **Changes in streamflow regimes and their response to different**
2 **soil and water conservation measures in Loess Plateau**
3 **watersheds, China**

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20 **Abstract**

21 Investigating the changes in streamflow regimes is useful for understanding the mechanisms
22 associated with hydrological processes in different watersheds and for providing information
23 to facilitate water resources management. In this study, we selected three watersheds, i.e.,
24 Sandu River, Hulu River, and Dali River on the Loess Plateau, to examine the changes in the
25 streamflow regimes and to determine their responses to different soil and water conservation
26 measures (terracing, afforestation, and damming). The daily runoff was collected
27 continuously by three hydrological gauges close to the outlets of the three watersheds from
28 1965 to 2016. The eco-surplus, eco-deficit, and degree of hydrological change were assessed
29 to detect hydrological alterations. The Budyko water balance equation was applied to estimate
30 the potential impacts of climate change and human activities on the hydrological regime
31 changes. Significant decreasing trends ($P < 0.05$) were detected in the annual streamflow in
32 the Sandu and Dali River watersheds, but not in the Hulu River watershed where afforestation
33 dominated. The annual eco-surplus levels were low and they decreased slightly at three
34 stations, whereas the eco-deficit exhibited dramatic increasing trends in the Sandu and Dali
35 River watersheds. In the Sandu River watershed (dominated by terraces), the runoff exhibited
36 the most significant reduction and the eco-deficit was the highest among the three watersheds.
37 The integral degrees of hydrological change were higher in the Sandu River watershed than
38 the other two watersheds, thereby suggesting substantial variations in the magnitude, duration,
39 frequency, timing, and rate of change in the daily streamflow. In the Dali River watershed
40 (dominated by damming), the changes in the extreme flow were characterized by a decreasing
41 number appearing in high flow. In these watersheds, human activities accounted for 74.1%

42 and 91.78% of the runoff reductions, respectively. In the Hulu River watershed (dominated by
43 afforestation), the annual runoff exhibited an insignificant decreasing trend but with a
44 significant increase in the low flow duration. Rainfall changes accounted for 64.30% of the
45 runoff reduction.

46 **Keywords:** climate variability; human activities; indicators of hydrologic alteration (IHA);
47 streamflow regime

48

49 **1. Introduction**

50 Streamflows are essential for environmental health, economic prosperity, and human well-
51 being. They provide the water required to produce energy, crops, and industrial products, and
52 for maintaining terrestrial and aquatic environment systems (Curmi et al., 2013). The water
53 resources in rivers determine the size, shape, structure, and dynamics of aquatic ecosystems.
54 The streamflow regime in a river is considered to be mainly responsible for the variations in
55 many other components of a river ecosystem, e.g., fish populations and nutrient cycling
56 (Richter et al., 2003). Streamflow variability influences hydrological functions and it is
57 important for maintaining biodiversity in rivers and the integrity of ecosystems (Arthington et
58 al., 2006; Jovanovic et al., 2016; Vogel et al., 2007).

59 Recent studies have demonstrated that streamflow regimes have exhibited obvious
60 changes in many rivers around the world. Approximately 24% of the world's large rivers
61 appear to have exhibited significant changes in the water flux according to observations of the
62 streamflow in 4399 rivers (Li et al., 2020). Forest clearing and large-scale agricultural
63 activities have led to dramatic changes in the streamflow in Amazonian rivers over the last 40
64 years (Aldrich et al., 2012; Latrubesse et al., 2009; Souza-Filho et al., 2016). Milly et al.
65 (2002) predicted increased streamflows for some areas of equatorial Africa, the La Plata
66 Basin, and high-latitude North America and Eurasia, but decreases in southern Africa,
67 southern Europe, the Middle East, and mid-latitude western North America by the year 2050.
68 The global mean annual temperature has increased by 0.8°C since 1880 (Flato et al., 2013),
69 and the humidity and precipitation increased by around 2% in the last century (Huntington,
70 2006; Wise, 2010), which may have altered the balance of the global water circulation and

71 energy cycle, thereby resulting in changes in the magnitude and spatiotemporal distribution of
72 streamflows (Nam et al., 2015). Climate change affects streamflow regimes by increasing
73 temperatures but also by altering precipitation patterns, rates of evaporation, transpiration, and
74 soil moisture contents (Merritt et al., 2006). Human activities may alter streamflow regimes
75 through land use changes, large-scale infrastructure such as reservoirs, water abstraction,
76 urbanization, and ecological restoration projects (Dey et al., 2017; Liang et al., 2015; Tamm et
77 al., 2018; Wang et al., 2016b). Both of these possible causes of changes in streamflow
78 regimes have been investigated widely (Li et al., 2021; Ellis, 2011; Guo et al., 2019; Munoz et
79 al., 2018; Zhang et al., 2017; Smith et al., 2016). It is necessary to explore streamflow
80 variations and related factors to better understand the mechanisms associated with streamflow
81 changes (Li and Fang, 2017; Wang et al., 2020), thereby providing useful reference data to
82 facilitate climate adaptation and improved water resources management.

83 Hydrologic alterations represent the changes in streamflow regimes at different temporal
84 and spatial scales due to regulation and water extraction via human activities. Over recent
85 decades, numerous methods have been developed to assess the streamflow variations in
86 rivers. The Newtonian and Darwinian approaches are currently applied widely from
87 microscopic and macroscopic perspectives, respectively (Omer et al., 2020; Wang et al.,
88 2020). Hydrological models (e.g., SWAT, TOPMODEL, and VIC) are mainly categorized as
89 Newtonian approaches, and they are useful tools for hydrologic alteration assessment and
90 decision making (Tamm et al., 2018). A representative Darwinian approach is the Budyko-
91 based method, which has been widely applied to quantify how climate change and human
92 activities might affect streamflow changes (Liang et al., 2015; Wang et al., 2019).

93 Many methods have been developed to assess streamflow regime changes. In particular,
94 Richter et al. (1996) proposed the indicators of hydrologic alteration (IHA) method for
95 quantifying streamflow regime changes using long-term observed daily flows in the Roanoke
96 River in North Carolina, USA. Furthermore, the range of variability approach (RVA) was
97 developed to meet the water requirements of ecosystems and achieve ecological sustainability,
98 where the degree to which the shape of a river's natural flow regime can be altered is a well-
99 established approach for quantifying flow regime alteration (Olden and Poff, 2003).

100 Many studies have demonstrated that hydrologic regimes are associated with dam
101 construction (Yang et al., 2008). Lai et al. (2013) and Wang et al. (2016b) found that the
102 operation of the Three Gorges Project aggravated hydrological droughts downstream, thereby
103 leading to extremely low water levels and environmental flow deficits, with substantial effects
104 on the streamflow regimes downstream of the Yangtze River. By contrast, Du et al. (2020)
105 suggested that the construction of dams can stabilize hydrological regimes and reduce the
106 flood peaks.

107 The Yellow River is one of China's largest rivers and the water it carries is utilized by
108 8% of the population. The lower reaches of the Yellow River appeared to have zero flows for
109 21 of the 27 years from 1972 to 1998, which contributed to the severe ecological damage to
110 the river (Chen et al., 2020). Droughts, water shortages, and soil and water losses are the most
111 severe environmental problems that have affected the social and economic development of the
112 Yellow River basin (Wu et al., 2004). The Loess Plateau is located in the middle reaches of
113 the Yellow River basin. Over 20 rivers flow into the Yellow River and they contribute nearly
114 9% of the sediment and 40% of the streamflow into the river basin. Due to severe soil erosion,

various soil and water conservation measures have been implemented on the Loess Plateau since the 1970s (Zhao et al., 2014). The main measures comprise afforestation, check dam construction, and terracing, which have increased rainfall infiltration and reduced the flood peaks, soil erosion, and sediment transported into the rivers. Many studies have been conducted on the Loess Plateau to estimate the changes in runoff and the sediment load, as well as their potential causes (Gao et al., 2015; Guo et al., 2020; Zhang et al., 2017; Zhang et al., 2018; Zhou et al., 2020). These studies have clarified the effects of climate changes and human activities on runoff and sediment load reductions. However, few studies have assessed the alterations in the streamflow regimes and their responses to different soil and water conservation measures. Due to the limited availability of information regarding conservation measures, few studies have also compared the hydrological responses of river flows to individual conservation measures. Therefore, the objectives of this study were: (1) to examine the changes in the streamflow regimes in three watersheds with different conservation measures; and (2) to quantify the effects of climate variability and human activities on the streamflow changes.

2. Study area and data

2.1. Study area

More than 20 tributaries in the middle reaches of the Yellow River flow through the Loess Plateau and they discharge approximately 90% of the sediment that enters the Yellow River. Since the 1950s, extensive soil and water conservation measures have been implemented in these watersheds to control severe soil erosion, including afforestation, grass

planting, terracing, and check dam construction. According to the spatial distribution and types of soil and water conservation measures on the Loess Plateau, we selected the following three different watersheds and analyzed the changes in their streamflow regimes: Sandu River watershed (dominated by terracing), Hulu River watershed (affected by afforestation), and Dali River watershed (dominated by check dams). Figure 1 shows the locations of the watersheds and hydro-meteorological gauges.

2.1.1. Sandu River watershed

The Sandu River watershed is located upstream of the Wei River basin and it covers an area of 2484 km². This watershed is characterized by a continental monsoon climate, with a mean annual temperature around 10.2°C and average annual precipitation of approximately 467.3 mm. The average annual streamflow and sediment yields are 0.45 billion m³ and 8560 t/km², respectively. Since the 1970s, many terraces have been constructed on the hill slopes. Terraces with a total area of 1807.45 km² were built by 2017, which covered 72.7% of the whole watershed.

2.1.2. Hulu River watershed

The Hulu River is a tributary of the Beiluo River (a tributary of the Wei River, Figure 1), with a drainage area of 4715 km². The mean annual precipitation and evaporation in this watershed are 494 mm and 1147 mm, respectively. The streamflow from July to September accounts for about 50% of the annual total. The catchment is covered by dense forest, where *Pinus tabuliformis* Carr. and *Platycladus orientalis* (Linn.) Franco are the major forest species present in this watershed. More than 90% of the sloping arable land was converted into forest or grassland after 1999 as part of the “Grain for Green” project launched by the Chinese

government. The vegetation cover has reached approximately 80% and the sediment yield decreased significantly from 107×10^4 t in the 1960s to 76.8×10^4 t in 2016.

2.1.3. Dali River watershed

The Dali River is the largest tributary of the Wuding River, with an area of 3906 km² and length of 170 km. The average annual streamflow was 1.31 billion m³ from 1960 to 2016. The average annual precipitation is 429 mm and high-intensity rainstorms cause considerable soil erosion in this watershed during the summer. Since the beginning of the 1980s, many check dams have been built in the watershed to trap sediment transported from hill slopes to the Yellow River. In 2017, 11602 check dams were present in the Wuding watershed, i.e., 1155 large sized check dams and 10174 medium to small sized check dams (Han et al., 2018). Due to the implementation of large-scale soil and water conservation measures in the Dali River watershed, the sediment load decreased from 0.64×10^8 t/a in the 1970s to 0.15×10^8 t/a from 2000 to 2016 (Zhang et al., 2019).

<Figure 1>

2.2. Data set

The daily streamflows at the Gangu (Sandu River), Zhangcunyi (Hulu River), and Suide (Dali River) hydrological stations (Figure 1) were obtained from the Hydrological Year Book of the Yellow River, which was published by the Ministry of Water Resources of China. In total, the daily discharge measurements for 52 years from 1965 to 2016 were collected for investigation. Table 1 shows general information for the three stations in the watersheds. Monthly rainfall, air temperature, relative humidity, sunshine duration, and wind speed

meteorological data at 11 climate stations were obtained from the China National Climate Center. The potential evapotranspiration was calculated using the Penman–Monteith equation (Allen et al., 1998). The basin-averaged precipitation and potential evapotranspiration data were estimated, and then interpolated with the inverse distance weighted method in ArcGIS 10.6 (<http://www.esri.com>). The spatial distributions of soil and water conservation measures, including terraces, check-dams, and reservoirs, in each catchment during 2017 were interpreted with Google Earth and validated based on a field survey. The homogeneity and continuity of all the measured data were checked to guarantee the data integrity and consistency before their release.

<Table 1>

3. Methodology

The Mann–Kendall test and accumulated anomaly method were applied to detect the changing trends and abrupt changes in the annual streamflow at the three hydrological stations. These methods have been used widely to examine changing hydro-meteorological time series (Kendall, 1975; Mann, 1945; Sagarika et al., 2014; Zhao et al., 2014), so they are not described in detail. The detailed estimation procedures were described in previous studies (Weber et al., 2010; Zhao et al., 2019). The IHA method (Richter et al., 1996) was employed to analyze the streamflow regime changes according to the daily streamflow, and the effects of climate change and human activities on streamflow variations were assessed by using the Budyko equation (He et al., 2019; Yang et al., 2014).

3.1. Changes in hydrological regimes

The IHA method was first developed by the Nature Conservancy and it has been applied widely. The IHA method employs 32 parameters (we excluded “the number of zero-flow days” because zero flows were never observed at the three hydrological stations) related to hydrological extremes and averages, which can be classified according to five major categories (Table 2).

The RVA method proposed by Richter et al. (1996) was used to quantify the hydrological changes in terms of the 32 indicators. The target range for the RVA comprises IHA values that fall within the thresholds between the 25th and 75th percentile values (Huang et al., 2019). The degree to which the RVA does not reach the target range denotes the degree of hydrological change, and can be expressed as (D_i) for each indicator. Cheng et al. (2019) and Zhang et al. (2016) described the calculation method in detail. Therefore, the following three

equal sized categories were employed: (1) $(|D_i| < 33\%)$ indicating little or no alteration; (2)

$(33\% < |D_i| < 67\%)$ denoting a moderate degree of alteration; and (3) $(67\% < |D_i| < 100\%)$

representing a large change.

The nondimensional eco-flow metrics comprising the eco-deficit and eco-surplus were estimated based on the flow duration curve (FDC) using the daily flow data (Vogel et al. 2007) in order to provide intuitive representations of the hydrological impacts and to supplement those characterized by the IHA. The annual FDC can be regarded as a function of the excess of the daily streamflow over the probability in each year. The eco-flow metrics

were estimated in the following four steps. (1) According to the change point, the daily streamflow series covering the period from 1965 to 2016 was subdivided into two sections. In the baseline period, human activities were regarded as having limited effects on the streamflow regimes. (2) The 25% and 75% FDC quantiles were taken as threshold values, and the values in the intervals were the adaptive range for the river ecosystem. (3) The eco-flow was determined by comparing the annual FDC in other years with the previously obtained 25% and 75% FDC quantiles. The area above the 75% quantile FDC was the eco-surplus, whereas the area below the 25% quantile FDC was the eco-deficit. Full details of the calculation procedure were provided by Gao et al. (2012) and Vogel et al. (2007).

<Table 2>

3.2. Impacts of climate variability and human activities

The Budyko equation is important in hydrology because it provides a concise and accurate representation of the relationship between the annual evapotranspiration and long-term average water and energy balance at catchment scales (Sposito, 2017). In a natural basin, the long-term average annual water and energy balance at the catchment scale can be expressed as follows.

$$R = P - E - \Delta S \quad (3)$$

The Budyko hypothesis (Budyko, 1974) considers the balance for precipitation (P) between potential evapotranspiration (E_0) and actual evapotranspiration (E). Based on the long-term catchment water balance equation, ΔS is assumed to be zero. By combining dimensional analysis and physical principles, Fu (1981) analytically derived the water–energy balance

function at the mean annual time scale, which is expressed as:

$$\frac{E}{P} = 1 + \frac{E_0}{P} - \left[1 + \left(\frac{E_0}{P} \right)^\omega \right]^{\frac{1}{\omega}}, \quad (4)$$

where E_0 is the mean annual potential evapotranspiration and the parameter ω represents the

catchment landscape characteristics. The long-term $\frac{E}{P}$ is mainly controlled by the water–

energy balance $\frac{E_0}{P}$ (the dryness index). Changes in the catchment streamflow can be

expressed as the sum of three components defined by Schaake (1990) as the precipitation,

potential evapotranspiration, and catchment landscape elasticity of the streamflow, and thus

the new equation is expressed as:

$$\Delta R = \varepsilon_P \frac{R}{P} \Delta P + \varepsilon_{E_0} \frac{R}{E_0} \Delta E_0 + \varepsilon_\omega \frac{R}{\omega} \Delta \omega = \Delta R_P + \Delta R_{E_0} + \Delta R_\omega, \quad (5)$$

where the elasticities of the streamflow are given as:

$$\varepsilon_P = \frac{\left[1 + \left(E_0 / P \right)^\omega \right]^{\frac{1}{\omega} + 1} - \left(E_0 / P \right)^{\omega + 1}}{\left[1 + \left(E_0 / P \right)^\omega \right] \left\{ \left[1 + \left(E_0 / P \right)^\omega \right]^{\frac{1}{\omega}} - \left(E_0 / P \right) \right\}} \quad (6)$$

$$\varepsilon_{E_0} = \frac{1}{\left[1 + \left(E_0 / P \right)^\omega \right] \left\{ 1 - \left[1 + \left(E_0 / P \right)^{-\omega} \right]^{\frac{1}{\omega}} \right\}} \quad (7)$$

$$\varepsilon_{\omega} = \frac{\ln [1 + (E_0 / P)^{\omega}] + (E_0)^{\omega} \ln [1 + (E_0 / P)^{-\omega}]}{\ln [1 + (E_0 / P)^{\omega}] \left\{ 1 - [1 + (E_0 / P)^{-\omega}]^{\frac{1}{\omega}} \right\}}, \quad (8)$$

where ω is a model parameter denoting the non-climatic effects on the water–energy balance attributed to the soil properties, topography, and vegetation (Gunkel et al., 2017; Wang et al., 2016a).

Changes in the observed mean annual runoff depth ΔR^T can be estimated between the baseline period and changing period, and they can be attributed to climate variability ΔR^C and human activities ΔR^h . The streamflow change due to climate variation (ΔR^C) includes the streamflow changes due to precipitation variation (ΔR_p) and potential evaporation variation (ΔR_{E_0}) (Koster & Suarez, 1999; Milly & Dunne, 2002). The contributions to annual streamflow changes due to climate change and human activities, respectively, can be approximated as follows.

$$\eta_C = \frac{\Delta R^C}{|\Delta R^C| + |\Delta R^h|} \times 100\% \quad (11)$$

$$\eta_h = \frac{\Delta R^h}{|\Delta R^C| + |\Delta R^h|} \times 100\% \quad (12)$$

4. Results

4.1. Changes in annual streamflows

4.1.1. Temporal variations in annual streamflows

Figure 2 shows the linear trends in the annual streamflows in the three watersheds. Overall, the annual streamflows tended to decrease at all stations, with relatively high reductions at Gangu station (0.159 mm/10a) and Suide station (0.138 mm/10a) during 1965–2016. For example, at Gangu station, the annual average streamflow was only $0.21 \times 10^8 \text{ m}^3/\text{a}$ from 1995 to 2016, which was much lower than that during 1965–1994 ($0.629 \times 10^8 \text{ m}^3/\text{a}$). The Mann–Kendall test showed that the annual streamflows at Gangu (Figure 2a) and Suide stations (Figure 2c) decreased significantly ($P < 0.01$), but not significantly at Zhangcunyi station (Figure 2b). Comparisons of the annual streamflow fluctuations at the three stations showed that the variability in the annual streamflow was higher at Zhangcunyi ($C_v = 0.34$).

<Figure 2>

4.1.2. Abrupt changes in annual streamflows

As shown in Figure 3, abrupt changes in the annual streamflows mostly occurred during the mid-1990s and they were mainly attributable to the large-scale soil and water conservation measures implemented in the middle reaches of the Yellow River. Abrupt changes were detected during 1994 at Gangu station (Figure 3a), 1990 at Zhangcunyi station (Figure 3b), and 1996 at Suide station (Figure 3c). Therefore, the total time series at each station were divided into two periods based on these breakpoints. The first period represented the baseline

period with very limited human activities or none. The second period represented the change period when the watershed experienced substantial changes in land use, afforestation/deforestation, and dam construction.

<Figure 3>

4.2. Changes in IHA metrics

The IHA indicators in group 1 (G1) represent the magnitudes of the average monthly median streamflow, which clearly varied at the three gauging stations. Significant decreases were found in all of the areas, but particularly at Gangu station, with average change rates lower than -50% ($P < 0.001$). The median 1-, 3-, 7-, 30-, and 90-day annual minimum/maximum flows (Group 2, G2) tended to decrease. The highest reduction in G2 was detected at Gangu station and the lowest at Zhangcunyi station. In addition, the 1- and 3-day minimum flows increased at Suide station but not significantly. The Group 3 (G3) indicators represent changes in the timing of extreme flows. The minimum and maximum dates tended to increase at all three stations, thereby suggesting time lag effects of the soil and water conservation measures. In particular, at Gangu station, the low flow date changed most greatly and reached up to 90.9%. In the Group 4 (G4), the durations of the low pulses tended to increase at all stations during the changing period, but especially at Gangu and Zhangcunyi stations, with higher change rates ($P < 0.05$) that increased from 3 to 6.5 days and 3.5 to 7.25 days, respectively. The durations of the high pulses remained relatively stable at all stations. Different trends were found at Suide station where the median duration of high pulses during

the change period decreased to 2.5 days and by 9.09% compared with that of 2.75 days before the baseline period. The flow fall rates all decreased significantly at three stations ($P < 0.05$).

The variations in the hydrologic indicators tended to differ among the stations, as shown in Figure 4b. We found that 13 indicators exhibited moderate alterations at Suide station, 18 indicators had low-degree alterations at Zhangcunyi station, and 75% of the indicators had high degree alterations at Gangu station. Clearly, the fall rates exhibited the greatest changes ($|Di| > 67\%$) in the three watersheds and they indicated declines in the streamflows, where more of the conditions in the post-period were below the lower limit of the RVA threshold than those in the pre-period. At Suide station, the Di value was related to the flood season. Indicators such as the duration of low pulses ($D_i = -78.57\%$), baseflow index ($D_i = -70\%$), and number of reversal stations were assigned to the high-degree alteration category at Zhangcunyi.

<Figure 4>

4.3. Changes in eco-flow metrics

In Figure 5, the blue and red curves correspond to the 25th percentile and 75th percentile FDC, respectively, during the baseline period at each station. Compared with the daily streamflow indices in the baseline period, the high and low flows decreased significantly at the three stations, where the reductions in the low flow (Q_{90}) rates were lower than those in the high flow (Q_{10}) rates at Gangu and Zhangcunyi stations. In particular, the low flow declined greatly by 69.57% at Gangu station. By contrast, the daily streamflow remained relatively stable at Zhangcunyi station, where the Q_{10} and Q_{90} components decreased by

23.48% and 18.4%, respectively. At Suide station, the reduction in the low flow component (44.64%) was higher than that in the high flow (23.33%) during the changing period.

<Figure 5>

Figure 6 shows the annual eco-flow metrics (eco-surplus and eco-deficit) obtained based on the annual FDC and temporal variations in the annual precipitation anomaly at the three stations. Overall, the annual eco-surplus and annual eco-deficit tended to fluctuate greatly. Remarkably, the variations in the annual eco-deficit were more substantial than those in the annual eco-surplus. Figures 6b, 6d, and 6e show the eco-surplus and eco-deficit results for the three stations in different decades. From the 1990s, the annual eco-deficit increased dramatically at Gangu station (a), but there was no apparent variation at Zhangcunyi (b) and only a slight increase at Suide station (c). The eco-surplus tended to decrease at the three stations. At Gangu station, persistent and high peaks were detected in the eco-deficit and eco-surplus during the early baseline period and later changing period. In addition, the negative deviation in the precipitation explained the eco-deficit during 1994–2002. At Zhangcunyi station, the eco-flow metrics varied consistently with the changes in precipitation, thereby suggesting that afforestation did not change the eco-flow metrics in the watershed and climate change may have contributed more to the streamflow changes. At Suide station, the eco-flow metrics and precipitation were strongly correlated. During the changing period, the eco-deficit increased because of the implementation of soil and water conservation measures, although the precipitation increased. According to Figure 6, the low flow rate contributed more to the

eco-surplus whereas the high flow rate contributed more to the eco-deficit. These results are similar to those reported by Du et al. (2020).

<Figure 6>

4.4. Attribution of streamflow variations to climate variability and human activities

Significant reductions in the annual streamflow could be attributed to climate change and human activities. We employed the Budyko equation to quantify the effects of climate change and land surface changes on the streamflow variations. After comparing the annual runoff rates in the changing period and baseline period, we found that the streamflow decreased by 16.83 mm, 4.16 mm, and 9.95 mm at Gangu, Zhangcunyi, and Suide stations, respectively. Thus, the annual streamflow reduction was lowest at Zhangcunyi station in the forest dominated watershed. As shown in Table 3, climate change only accounted for 25.9% and 8.22% of the reductions in the Sandu and Dali River watersheds, respectively. Human activities were mainly responsible for the runoff reductions, particularly in the Dali River watershed (91.78%). In contrast to the Sandu and Dali River watersheds, the streamflow reduction in the Hulu River basin was attributed primarily to climate change (64.30%), and human activities (mainly afforestation) were only responsible for the other 35.70%.

<Table3>

5. Discussion

5.1. Impacts of climate change and human activities on streamflows

Using hydro-climatic data, the elasticity/sensitivity method based on the Budyko equation has been widely applied to quantify the effects of climate variability and human activities on streamflow changes. Wang et al. (2020) found that human activities accounted for over 50% of the runoff reductions in four catchments in the Yellow River basin from 1960 to 2015. Gu et al. (2019) showed that precipitation and human activities contributed approximately 20% and 80%, respectively, to the runoff reductions in both river sources and the middle reaches of the Yellow River basin. Similar results were obtained in our analyses of watersheds dominated by terraces and check dams, but different results in the afforested watershed (Hulu River watershed).

To further verify our results, we applied the double mass curve method to quantify the effects of climate change and human activities on the runoff changes (Chang et al., 2015; Gao et al., 2017). Figure 7 shows that changes were found in the cumulative annual runoff curves and precipitation at Suide and Gangu stations, thereby indicating that the runoff changes were more significant in these two watersheds. The estimates indicate that human activities accounted for 86.85% of the runoff reduction at Gangu station and 97.11% at Suide station. For the Hulu River watershed, human activities accounted for 35.70% of the streamflow reduction and the remaining 64.30% was attributed to climate change. Thus, the results obtained using the double mass curve were consistent with those produced by the Budyko method.

<Figure 7>

5.2. Impacts of soil and water conservation measures on streamflow regime changes

Soil and water conservation measures have been implemented in the upper and middle reaches of the Yellow River basin since the late 1950s (Zhang et al., 2018). These measures include biological measures comprising afforestation and grassing, and engineering measures comprising terracing and check dams. Approximately 58000 check dams have been built, including 5546 large sized dams. The area covered by terraces is 5.5×10^4 km², mostly upstream of the Wei River and in the middle reaches of the Yellow River basin.

The proportion of vegetation cover in the study area was 28.6% in the 1980s and it increased to 63.2% by 2018. These obvious changes in the land surface cover were responsible for the significantly reduced river streamflow. Vegetation plays a vital role in regulating terrestrial water flows, where forests can fix and store carbon as well as regulating water functionalities (Bai et al., 2020; Ellison et al., 2017; Farooqi et al., 2020). The “Grain for Green” project launched in 1999 greatly increased the vegetation cover and reduced soil erosion on the Loess Plateau (Zhou et al., 2015). Figure 8 shows the vegetation cover at different levels (high, medium, and low vegetation cover) in the three watersheds. Changes in the medium and high vegetation cover levels occurred in the three watersheds. In general, the Hulu River watershed had a high vegetation cover rate, and transformations occurred from medium vegetation to high vegetation cover in the other two watersheds from 1998 to 2016 (Figures 8b and 8c). These results are consistent with previous reports of great increases in the

vegetation cover in the middle reaches of the Yellow River basin (Wang et al., 2019), and the vegetation cover has increased significantly since 2002 (Xin et al. 2008).

<Figure 8>

The hydrological responses to afforestation differed among the three river basins. We found significant reductions in the runoff in the Sandu and Dali River watersheds, and a constant increase in the eco-deficit values after the abrupt change. The results suggest that engineering measures had more significant effects on runoff reduction. A non-significant decrease in the annual runoff was found in the forest-dominated basin (Hulu River basin), thereby confirming that afforestation affected the river streamflow by increasing terrestrial interception and evapotranspiration (*ET*). Moreover, large differences in the eco-flow metrics and degree of hydrologic alteration according to most of the IHA indicators were found between the Hulu River basin and other watersheds. Previous studies by Zhou et al. (2015), Ellison et al. (2017), and Evaristo et al. (2019) showed that afforestation can effectively moderate floods by storing or recycling substantial amounts of water via interception, infiltration, transpiration, evaporation, and groundwater recharge. Vegetation cover can help to maintain a low flow by moderating the streamflow and conserving water. The change in precipitation was also an important factor that affected the variations in vegetation cover. Thus, the relatively stable streamflow regime in the Hulu River basin was caused by afforestation.

However, compared with the changes in the annual runoff in the humid region, the dry region tended to decrease due to both climate variability and afforestation. Zhou et al. (2020)

reported that the water conserving effects of afforestation were higher than those of increased terrestrial interception and evapotranspiration in a humid region. However, this is the opposite of the effect found in dry areas, such as the Hulu River watershed. Afforestation can increase low flows and reduce high floods through more interception and infiltration, but it also results in higher evapotranspiration. This may explain the decreased runoff at Zhangcunyi station.

<Figure 9>

Terraced fields are essential measures for reducing the transport of soil eroded from upstream to downstream areas by reshaping the microtopography and increasing the slope length (Tarolli et al., 2014), as demonstrated by Zhang et al. (2008) and Chen et al. (2017). An experimental study conducted by Ran et al. (2006) suggested that terraces could reduce the runoff by 60.7% and sediment yield by 58.0% in the Wei River Basin. In the Sandu River basin, about 72.7% of the watershed is covered by terraces (Table 4), which is much higher than the coverage rates found in the Dali and Hulu River basins. The change in the streamflow regime in the Sandu River basin differed from those in the other watersheds. The annual streamflow and eco-deficit decreased most in the Sandu River basin watershed, thereby demonstrating the significant effect of terraces on the surface runoff. The Budyko method indicated that the evaporation was 40 mm higher during the changing period than the baseline period. Moreover, the duration of the low flow increased, whereas the number of high flows clearly decreased.

<Table 4>

473

474 Check dams were initially built to trap sediment and produce fertile agricultural land in
475 the gully-dominated region of the Loess Plateau. According to Li et al. (2016), check dams
476 are vital engineering measures for retaining flood water, trapping upstream sediment,
477 increasing the availability of farmland for agricultural production, and reducing downstream
478 sediment transport. In the Dali River watershed, 46.7% of the total area was controlled by
479 check dams and reservoirs in 2017. By contrast, far fewer check dams were present in the
480 other two watersheds. Dams and reservoirs can effectively reduce flood peaks and increase
481 their slow time, as indicated by the IHA metrics. Dam construction significantly increases the
482 seepage of soil water into groundwater, thereby leading to an increase in low flows.
483 According to Martin-Rosales et al. (2007), check dams increased the infiltration of runoff by
484 3–50% in a semiarid region of Spain. Unlike vegetation measures, engineering measures can
485 immediately regulate surface runoff, especially in high flow events. Furthermore, the
486 hydrological alteration degree in the Dali River watershed changed significantly in terms of
487 the streamflow regime among the watersheds, thereby demonstrating the strong influence of
488 dams on hydrological processes.

489 Previous studies employed the IHA/RVA approach to analyze the hydrological
490 alterations in the middle reaches of the Yellow River. Zhang et al. (2016) determined the
491 effects of reservoir construction and operation in the upper Yellow River basin on alterations
492 in the ecological flow regimes, and found decreases in high flows and increases in low flows
493 at mainstream stations. Indicators of low flow and rising and falling water conditions changed
494 greatly, and they were consistent with our results. However, we obtained more detailed results

regarding the streamflow regime changes related to different soil and water conservation measures.

6. Conclusion

In this study, we applied the IHA/RVA method and eco-flow metrics to estimate streamflow regimes based on the daily discharge in three watersheds where different soil and water conservation measures were implemented from 1965 to 2016 on the Loess Plateau. The Budyko equation was employed to quantify the effects of climate change and human activities on the streamflow variations. Our main conclusions can be summarized as follows.

(1) The annual streamflows decreased significantly at Gangu station ($Z = -6.75$) in the Sandu River watershed and Suide station ($Z = -3.28$) in the Dali River watershed. The decrease in the annual streamflow was not significant at Zhuangcunyi station ($Z = -1.75$) in the Hulu River watershed. Abrupt change points were mostly detected in the 1990s for all watersheds.

(2) The hydrological indicators obtained with the IHA/RVA method showed that the degrees of hydrological change in the Sandu River, Dali River, and Hulu River basins were at high, medium, and low levels, respectively, thereby suggesting that the streamflow regimes varied under different soil conservation measures. On the microscale level, the low flow duration and the number of high flows had great influences at the three stations.

(3) The changes in the eco-deficit were more remarkable than those in the eco-surplus at all stations. The annual eco-deficit increased dramatically after the 1980s at Gangu station, but there was no obvious change at Zhangcunyi and it only increased moderately after the

1990s at Suide station. The annual eco-surplus tended to decrease slightly at all three stations.

(4) The significant annual runoff reductions were primarily attributed to human activities in the Dali River watershed (91.7%) and Sandu River watershed (74.1%). Climate change accounted for 64.30% of the annual streamflow decrease in the Hulu River watershed where afforestation dominated.

The streamflow at Gangu station changed greatly due to the significant effect of numerous terraces, which apparently led to an eco-deficit. The flow regime was altered in the forest dominated watershed but the overall streamflow was relatively weaker and vulnerable to climate change. Moreover, human activities made the greatest contribution to the reduced runoff in the Suide watershed. The results obtained in this study provide novel insights into hydrological regime changes and their responses to different soil and water conservation measures.

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Data Availability Statement

Data available on request due to privacy/ethical restrictions.

539 **References**

- 540 Aldrich, S., Walker, R., Simmons, C., Caldas, M., Perz, S., 2012. Contentious Land Change in the
541 Amazon's Arc of Deforestation. *Annals of the Association of American Geographers* 102,
542 103–128.
- 543 Allen, R. G., Pereira, L. S., Raes, D., Smith, M., 1998. *FAO Irrigation and Drainage Paper No. 56*.
544 Rome: Food and Agriculture Organization of the United Nations 56, 97–156.
- 545 Arthington, A.H., Bunn, S.E., Poff, N.L., Naiman, R.J., 2006. The challenge of providing
546 environmental flow rules to sustain river ecosystems. *Ecological Applications* 16, 1311–
547 1318.
- 548 Bai, P., Liu, X.M., Zhang, Y.Q., Liu, C.M., 2020. Assessing the impacts of vegetation greenness
549 change on evapotranspiration and water yield in China. *Water Resources Research* 56.
- 550 Budyko, M. I. 1974. *Climate and Life*. International Geophysics Series 18, 294.
- 551 Chang, J.X., Wang, Y.M., Istanbuluoglu, E., Bai, T., Huang, Q., Yang, D.W., Huang, S.Z., 2015.
552 Impact of climate change and human activities on runoff in the Weihe River Basin, China.
553 *Quaternary International* 380–381, 169–179.
- 554 Chen, Y.P., Fu, B.J., Zhao, Y., Wang, K.B., Wang, H., 2020. Sustainable development in the
555 Yellow River Basin: Issues and strategies. *Journal of Cleaner Production* 263, 121223.
- 556 Cheng, J.X., Xu, L.G., Fan, H.X., Jiang, J.H., 2019. Changes in the flow regimes associated with
557 climate change and human activities in the Yangtze River. *River Research and Applications*
558 35, 1415–1427.
- 559 Chen, T., Christensen, M., Nan, Z.B., Hou, F.J., 2017. The effects of different intensities of long-
560 term grazing on the direction and strength of plant-soil feedback in a semiarid grassland of

561 Northwest China. *Plant Soil* 413, 303–317.

562 Curmi, E., Richards, K., Fenner, R., Allwood, J.M., Kopec, G.M., Bajzelj, B., 2013. An integrated
563 representation of the services provided by global water resources. *Journal of Environmental*
564 *Management* 129, 456–462.

565 Dey, P., Mishra, A., 2017. Separating the impacts of climate change and human activities on
566 streamflow: A review of methodologies and critical assumptions. *Journal of Hydrology* 548,
567 278–290.

568 Du, J., Wu, X., Wang, Z., Li, J., Chen, X., 2020. Reservoir-induced hydrological alterations using
569 ecologically related hydrologic metrics: case study in the Beijiang River, China. *Water* 12, 1–
570 19.

571 Ellis, E., C., 2011. Anthropogenic transformation of the terrestrial biosphere. *Philosophical*
572 *Transactions of the Royal Society A Mathematical Physical and Engineering Sciences* 369,
573 1010–1035.

574 Ellison, D., Morris, C.E., Locatelli, B., Sheil, D., Cohen, J., Murdiyarso, D., Gutierrez, V.,
575 Noordwijk, M.V., Creed, I.F., Pokorny, J., 2017. Trees, forests and water: Cool insights for a
576 hot world. *Global Environmental Change* 43, 51–61.

577 Evaristo, J., McDonnell, J.J., 2019. Global analysis of streamflow response to forest management.
578 *Nature* 574, 455–461.

579 Farooqi, T., Li, X., Yu, Z., Liu, S., Sun, O.J., 2020. Reconciliation of research on forest carbon
580 sequestration and water conservation. *Journal of Forestry Research* 32, 7–14.

581 Flato, G., Marotzke, J., Abiodun, B., Braconnot, P., Chou, S.C., Collins, W.J., Cox, P., Driouech,
582 F., Emori, S., Eyring, V., 2013. Evaluation of Climate Models. In: *Climate Change 2013: The*
583 *Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of*

584 the Intergovernmental Panel on Climate Change. *Computational Geometry* 18(2), 95–123.

585 Fu, B. P. 1981. On the calculation of the evaporation from land surface. *Chinese Journal of*

586 *Atmospheric Sciences*, 5, 23–31.

587 Gao, B., Yang, D., Zhao, T.T.G., Yang, H.B, 2012. Changes in the eco-flow metrics of the Upper

588 Yangtze River from 1961 to 2008. *Journal of Hydrology* 448–449, 30–38.

589 Gao, P., Jiang, G., Wei, Y., Mu, X., Wang, F., Zhao, G., Sun, W., 2015. Streamflow regimes of the

590 Yanhe River under climate and land use change, Loess Plateau, China. *Hydrological*

591 *Processes* 29, 2402–2413.

592 Gao, P., Li, P.F., Zhao, B.L., Xu, R.R., Zhao, G.J., Sun, W.Y., Mu, X.M., 2017. Use of double mass

593 curves in hydrologic benefit evaluations. *Hydrological Processes* 31, 4639–4646.

594 Gu, C.J., Mu, X.M., Gao, P., Zhao, G. J., Yu, Q., 2019. Rainfall erosivity and sediment load over

595 the Poyang Lake basin under variable climate and human activities since the 1960s.

596 *Theoretical and Applied Climatology* 136, 1–16.

597 Gunkel, A., Lange, J., 2017. Water scarcity, data scarcity and the Budyko curve - An application in

598 the Lower Jordan River Basin. *Journal of Hydrology Regional Studies* 12, 136–149.

599 Guo, Y., Fang, G., Xu, Y.P., Tian, X., Xie, J., 2019. Identifying how future climate and land

600 use/cover changes impact streamflow in Xinanjiang Basin, East China. *Science of the Total*

601 *Environment* 710, 136275.

602 Han, X. N., Xie, S.Y., Gao, Y.F., 2018. Study on sediment retaining effect of check dams in

603 Wuding River basin in recent years. *Yellow River* 40(11), 5–8.

604 He, Y., Song, J.X., Hu, Y.Y., Tu, X., Zhao, Y., 2019. Impacts of different weather conditions and

605 landuse change on runoff variations in the Beiluo River Watershed, China. *Sustainable Cities*

606 *and Society* 50, 101674.

607 Huang, Y.H., Huang, B.B., Qin, T.L., Nie, H.J., Shen, Z.Q., 2019. Assessment of hydrological
608 changes and their influence on the aquatic ecology over the last 58 years in Ganjiang Basin,
609 China. *Sustainability* 11, 4882.

610 Huntington, T., G., 2006. Evidence for intensification of the global water cycle: Review and
611 synthesis. *Journal of Hydrology* 319(13), 83–95.

612 Jovanovic, T., García, S., Gall, H., Mejía, A., 2016. Complexity as a streamflow metric of
613 hydrologic alteration. *Stochastic Environmental Research and Risk Assessment* 31(8), 2107–
614 2119.

615 Kendall, M.G., 1990. Rank correlation methods. *British Journal of Psychology* 25, 86–91.

616 Koster, R.D., Suarez, M.J., 1999. A simple framework for examining the interannual variability of
617 land surface moisture fluxes. *Journal of Climate* 12, 1911–1917.

618 Lai, X., Jiang, J., Yang, G., Lu, X.X., 2013. Should the Three Gorges Dam be blamed for the
619 extremely low water levels in the middle-lower Yangtze River? *Hydrological Processes* 28.
620 150–160.

621 Latrubesse, E.M., Amsler, M.L., Morais, R.P.D., Aquino, S., 2009. The geomorphologic response
622 of a large pristine alluvial river to tremendous deforestation in the South American tropics:
623 The case of the Araguaia River. *Geomorphology* 113, 239–252.

624 Li, E.H., Mu, X. M., Zhao, G. J., Gao, P., Sun, W. Y., 2016. Effects of check dams on runoff and
625 sediment load in a semi-arid river basin of the yellow river. *Stochastic Environmental*
626 *Research and Risk Assessment* 31, 1791–1803.

627 Li, L., Ni, J.R., Fang, C., Yao, Y., Frolova, N., Magritsky, D., Borthwick A.G.L., Ciais, P., Wang,
628 Y.C., Zheng, C.M., Walling, D.E. 2020. Global trends in water and sediment fluxes of the
629 world's large rivers. *Science Bulletin* 65, 62–69.

630 Li, J., Xia, J.Q., Ji,Q.F., 2021. Rapid and long-distance channel incision in the Lower Yellow River
631 owing to upstream damming. *Catena* 196, 104943.

632 Liang, W., Bai, D., Wang, F., Fu, B., Yan, J., Wang, S., Yang, Y., Long, D., Feng, M., 2015.
633 Quantifying the impacts of climate change and ecological restoration on streamflow changes
634 based on a Budyko hydrological model in China's Loess Plateau. *Water Resources Research*
635 51(8), 6500–6519.

636 Milly, P. C. D., Dunne, K. A., 2002. Macroscale water fluxes 2. Water and energy supply control
637 of their interannual variability. *Water Resources Research* 38(24), 1–9.

638 Mann, H.B., 1945. Nonparametric test against trend. *Econometrica* 13, 245–259.

639 Martin-Rosales, W., Gisbert, J., Pulido-Bosch, A., Vallejos, A., Fernandez-Cortes, A., 2007.
640 Estimating groundwater recharge induced by engineering systems in a semiarid area
641 (southeastern Spain). *Environmental Geology* 52, 985–995.

642 Merritt, W.S., Alila, Y., Barton, M., Taylor, B., Cohen, S., Neilsen, D., 2006. Hydrologic response
643 to scenarios of climate change in sub watersheds of the Okanagan basin, British Columbia.
644 *Journal of Hydrology* 326, 79–108.

645 Nam, W.H., Choi, J.Y., Hong, E.M., 2015. Irrigation vulnerability assessment on agricultural water
646 supply risk for adaptive management of climate change in South Korea. *Agricultural Water*
647 *Management* 152, 173–187.

648 Munoz, S. E., Giosan, L., Therrell, M. R., Jonathan,W. F., Remo, Z. S., Shen ,Z. X., Sullivan,
649 R.M., Wiman, C., Donnel, M.O., Dnneal, J.P. 2018. Climatic control of Mississippi River
650 flood hazard amplified by river engineering. *Nature* 556, 95–98.

651 Olden, J.D., Poff, N.L., 2010. Redundancy and the choice of hydrologic indices for characterizing
652 streamflow regimes. *River Research and Applications* 19, 101–121.

653 Omer, A., Ma, Z.G., Zheng, Z.Y., Saleembc, F. 2020. Natural and anthropogenic influences on the
654 recent droughts in Yellow River Basin, China. *Science of the Total Environment* 704,135428.

655 Ran, D.C., 2006. Water and sediment variation and ecological protection measures in the middle
656 reach of the Yellow River. *Resource Science* 28, 93–100.

657 Richter, B. D., Baumgartner, J. V., Powell, J., Braun, D.P., 1996. A method for assessing
658 hydrologic alteration within ecosystems. *Conservation Biology* 10, 1163–1174.

659 Richter, B.D., Mathews, R., Wigington, H.R., 2003. Ecologically sustainable water management:
660 Managing river flows for ecological integrity. *Ecological Applications* 13, 206–224.

661 Sagarika, S., Kalra, A., Ahmad, S., 2014. Evaluating the effect of persistence on long-term trends
662 and analyzing step changes in streamflow of the continental United States. *Journal of*
663 *Hydrology* 517, 36–53.

664 Schaake, J.C., 1990. From climate to flow. In: Waggoner PE (ed.) *Climate change and US water*
665 *resources*. John Wiley, New York, 177–206.

666 Smith, N.D., Morozova, G.S., Pérez-Arlucea, M., Gibling, M.R., 2016. Dam-induced and natural
667 channel changes in the Saskatchewan River below the E.B. Campbell Dam, Canada.
668 *Geomorphology* 269,186–202.

669 Souza-Filho, P.W.M., Souza, E.B.D., Júnior, R.O.S., Nascimento, W.R., Mendonça, B.R.V.D.,
670 Guimarães, J.T.F., Dall’Agnol, R., Siqueira, J.O., 2016. Four decades of land-cover, land-use
671 and hydroclimatology changes in the Itacaiúnas River watershed, southeastern Amazon.
672 *Journal of Environmental Management* 167, 175–184.

673 Sposito,G., 2017. Understanding the Budyko Equation. *Water* 9, 236.

674 Tamm, O., Maasikame, S., Padari, A., Tamm, T., 2018. Modelling the effects of land use and
675 climate change on the water resources in the eastern Baltic Sea region using the SWAT

676 model. *Catena* 167, 78–89.

677 Tarolli, P., Preti, F., Romano, N., 2014. Terraced landscapes: from an old best practice to a
678 potential hazard for soil degradation due to land abandonment. *Anthropocene* 6, 10–25.

679 Vogel, R.M., Sieber, J., Archfield, S.A., Smith, M.P., Apse, C.D., Huber-Lee, A., 2007. Relations
680 among storage, yield, and instream flow. *Water Resources Research* 43, 909–918.

681 Wang, C., Wang, S., Fu, B., Zhang, L., 2016a. Advances in hydrological modelling with the
682 Budyko framework: A review. *Progress in Physical Geography* 40(3), 409–430.

683 Wang, F., Duan, K., Fu, S., Gou, F., Liang, W., Yan, J., Zhang, W., 2019. Partitioning climate and
684 human contributions to changes in mean annual streamflow based on the Budyko
685 complementary relationship in the Loess Plateau, China. *Science of the Total Environment*
686 665, 579–590.

687 Wang, W., Zhang, Y. Y., Tang, Q. H. 2020. Impact assessment of climate change and human
688 activities on streamflow signatures in the Yellow River Basin using the Budyko hypothesis
689 and derived differential equation. *Journal of Hydrology* 591, 125460.

690 Wang, W., Sun, L., Luo, Y., 2019. Changes in vegetation greenness in the upper and middle
691 reaches of the Yellow River Basin over 2000–2015. *Sustainability* 11, 2176.

692 Wang, Y.K., Rhoads, B.L., Wang, D., 2016b. Assessment of the flow regime alterations in the
693 middle reach of the Yangtze River associated with dam construction: potential ecological
694 implications. *Hydrological Processes* 30, 3949–3966.

695 Weber, K., Stewart, M., 2010. A critical analysis of the cumulative rainfall departure concept.
696 *Ground Water* 42, 935–938.

697 Wise, E.K., 2010. Climate–streamflow linkages in the North-Central Rocky Mountains:
698 implications for a changing climate. *Annals of the Association of American Geographers* 100,

699 806–817.

700 Wu, F.Q., 2004. Study on the benefits of level terrace on soil and water conservation. *Science of*

701 *Soil and Water Conservation* 2, 34–37.

702 Yang, H., Yang, D., Lei, Z., Sun, F., 2008. New analytical derivation of the mean annual water-

703 energy balance equation. *Water Resources Research* 44, 893–897.

704 Yang, H.B., Qi, J., Xu, X. Y., Yang, D.W., Lv, H.F., 2014. The regional variation in climate

705 elasticity and climate contribution to runoff across China. *Journal of Hydrology* 517, 607–

706 616.

707 Zhang, H.B., Singh, V.P., Zhang, Q., Gu, L., Sun, W.B., 2016. Variation in ecological flow regimes

708 and their response to dams in the upper Yellow River basin. *Environmental Earth Sciences*

709 75, 1–16.

710 Zhang, K.L., Shu, A.P., Xu, X.L., Yang, Q.K., Yu, B., 2008. Soil erodibility and its estimation for

711 agricultural soils in China. *Journal of Arid Environments* 72, 1002–1011.

712 Zhang, L., Karthikeyan, R., Bai, Z., Srinivasan, R., 2017. Analysis of streamflow responses to

713 climate variability and land use change in the Loess Plateau region of China. *Catena* 154, 1–

714 11.

715 Zhang, X., Lin, P., Chen, H., Yan, R., Zhang, J., Yu, Y., Liu, E., Yang, Y., Zhao, W., Lv, D., 2018.

716 Understanding land use and cover change impacts on run-off and sediment load at flood

717 events on the Loess Plateau, China. *Hydrological Processes* 32, 576–589.

718 Zhang, Y., Bi, Z., Zhang, X., Yu, Y., 2019. Influence of landscape pattern changes on runoff and

719 sediment in the Dali River Watershed on the Loess Plateau of China. *Land* 8(12), 1–12.

720 Zhao, G. J., Tian, P., Mu, X.M., Jiao, J., Wang, F., Gao, P., 2014. Quantifying the impact of climate

721 variability and human activities on streamflow in the middle reaches of the Yellow River

722 basin, China. *Journal of Hydrology* 519, 387–398.

723 Zhao, Q.Q., Wang, L., Liu, H., Zhang, Q., 2019. Runoff and sediment variation and attribution

724 over 60 years in typical Loess Plateau basins. *Journal of Soils and Sediments* 19(10), 3631–

725 3647.

726 Zhou, G.Y., Wei, X.H., Chen, X.Z., Zhou, P., Liu, X.D., Xiao, Y., Sun, G., Scott, D.F., Zhou,

727 S.Y.D., Han, L.S., Sun, Y.X., 2015. Global pattern for the effect of climate and land cover on

728 water yield. *Nature Communications* 6, 5918.

729 Zhou, X., Huang, X., Zhao, H., Ma, K., 2020. Development of a revised method for indicators of

730 hydrologic alteration for analyzing the cumulative impacts of cascading reservoirs on flow

731 regime. *Hydrology and Earth System Sciences* 24, 4091–4107.