

# 1 Response of soil aggregate disintegration to the different content of 2 organic carbon and its fractions during splash erosion

## 3 Abstract

4 Aggregate disintegration is a critical process in soil splash erosion. However,  
5 the effect of soil organic carbon (SOC) and its fractions on soil aggregates  
6 disintegration is still not clear. In this study, five soils with similar physical and  
7 chemical properties and different contents of SOC have been used. The effects of  
8 slaking and mechanical striking on splash erosion were distinguished by using  
9 deionized water and 95% ethanol as raindrops. The simulated rainfall experiments  
10 were carried out in four heights (0.5, 1.0, 1.5, and 2.0 m). The result indicated that  
11 the soil aggregate stability increased with the increases of SOC and light fraction  
12 organic carbon (LFOC). The relative slaking and the mechanical striking index  
13 increased with the decreases of SOC and LFOC. The reduction of macroaggregates  
14 in eroded soil gradually decreased with the increase of SOC and LFOC, especially in  
15 alcohol test. The amount of macroaggregates ( $>0.25\text{mm}$ ) in deionized water tests  
16 were significantly less than that in alcohol tests under the same rainfall heights. The  
17 contribution of slaking to splash erosion increased with the decrease of heavy  
18 fractions organic carbon (HFOC). The contribution of mechanical striking was  
19 dominant when the rainfall kinetic energy increased to a range of threshold between  
20  $9\text{ J m}^{-2}\text{ mm}^{-1}$  and  $12\text{ J m}^{-2}\text{ mm}^{-1}$ . This study could provide the scientific basis for deeply  
21 understanding the mechanism of soil aggregates disintegration and splash erosion.

## 22 1. Introduction

23 Splash erosion is an initial stage and an important component of interrill erosion  
24 (Kinnell, 2005; Van Dijk, Meesters, & Bruijnzeel, 2002). The aggregate  
25 fragmentation caused by raindrops striking is the first critical process in splash

26erosion (Legout, Leguedois, Le Bissonnais, & Lssa, 2005; Shainberg, Levy,  
27Rengasamy, & Frenkel, 1992). Many studies have showed that raindrops striking can  
28damage the soil structure, disperse and transport soil particles, thereby reducing soil  
29permeability. The amount of soil particles splashed increases with the speed at which  
30the raindrop hits the ground (Moss & Green, 1987). Fu et al. (2017) found that there  
31are significant exponential relations between the distance of splash and the size of  
32raindrops. Xiao et al. (2017) reported that the contribution of slaking decreasing with  
33the increased of the rainfall kinetic energy in splash, and that the contribution of  
34mechanical striking was opposite. It was also found that the soil texture plays vital  
35role in soil aggregates disintegration during splash erosion when the soil organic  
36carbon (SOC) was low (Xiao et al., 2018).

37 Soil aggregate stability is a crucial physical indicator that determine  
38disintegration resistance in soil erosion process, it determines the resistance of the  
39soil to erosion (Barthes & Roose, 2002; Bronick & Lal, 2005; Nichols & Toro,  
402011). The SOC plays an important role in formation of aggregates (Reeves, 1997;  
41Smith & Petersen, 2000; Tisdall & Oades, 1982). Furthermore, the physical  
42protection of aggregates is one of the main stabilizing mechanisms of SOC (Feller &  
43Beare, 1997; Li & Pang, 2014; Tisdall & Oades, 1982). Emerson (1967) found that  
4490% of SOC in the topsoil was in aggregates. Six et al. (1998; 2004) reported that  
45the SOC is one of the most important cementing materials in aggregates. Tisdall and  
46Oades (1982) proposed the soil aggregate theoretical model, which states that  
47microaggregate is an important precondition to form macroaggregate.

48 In order to distinguish the influence of SOC on the formation of aggregates,  
49some researchers used physical or chemical methods to classify SOC into different  
50components (Barrios, Buresh, & Sprent, 1996; Dhillon & Van Rees, 2017; Guan et  
51al, 2018; Six, Paustian, Elliott, & Combrink, 2000; Xiang, Zhang, & Wen, 2015).

52 Density fractionation of SOC is a common way to separate the SOC into light  
53 fraction organic carbon (LFOC) and heavy fraction organic carbon (HFOC)  
54 (Christensen, 1992; Turchenek & Oades, 1979; Wagai, 2009). The LFOC was labile  
55 fractions that represent an intermediate organic carbon pool between humified  
56 organic matter and undecomposed residues (Janzen, Campbell, Brandt, Lafond, &  
57 Townley Smith, 1992). The HFOC was stable that has lower carbon concentrations  
58 and slow decomposition rate and transformation rate (Golchin, Clarke, Oades, &  
59 Skjemstad, 1995a; Golchin & Oades, 1995b; Hassink, 1995; John, Yamashita,  
60 Ludwig, & Flessa, 2005). Oades (1984) and Elliott (1986) found that the roots and  
61 the fungal hyphae in LFOC can promote the formation of aggregates directly. These  
62 studies focused on the formation process of aggregates or the influences of different  
63 components of organic carbon on the particle size of aggregates (Guan et al., 2018;  
64 Holeppass, Singh, & Lal, 2004; Xiang et al., 2015). However, the role of SOC and its  
65 fractions in inhibiting the destruction of aggregates during splash erosion is still not  
66 clear.

67 The purposes of this study were (i) to explore the influence of SOC and its  
68 fractions on the disintegration of aggregates; (ii) to quantify the effects of SOC on  
69 contribution of slaking and mechanical striking to splash erosion.

## 70 2. Materials and methods

### 71 2.1 Study area and sampling

72 Five soils were collected from Fuxian (36°03'~36°04'N, 109°08'~109°09'E) in  
73 Shaanxi province, China. The soils with distinct SOC contents, due to different years  
74 after conversion from cropland to forestry, were selected. The sampling areas belong  
75 to hilly and gully regions of the Loess Plateau and the main soil type is loessial soil.  
76 The average annual temperature and precipitation in this area are 9°C and 576 mm,  
77 respectively. The main types of land use were forest. The vegetation was mainly

78 *Ulmus pumila*, *Betula platyphlla*, *Prunus*, *Populus davidiana* and so on.

79 Soil samples were collected from the uppermost 20 cm layer, air-dried and  
80 sieved through a 5 mm sieve to remove roots and gravels. The basic physical-  
81 chemical properties of the soils in the experiment are showed in Table 1.

## 822.2 Experimental design

83 The rainfall device is composed of three parts, including rainfall liquid supply,  
84 raindrop generation and supporting frame (Figure 1). The raindrop generation part  
85 was a cylinder with diameter of 30 cm which is made from smooth steel. Thirty-nine  
86 rainfall needles with a diameter of 0.6 mm were evenly distributed on the bottom of  
87 the cylinder. The fall height was controlled by adjusting the height of the supporting  
88 frame. Splash pan is an inverted cone device with an outer ring (30 cm diameter on  
89 top, 10 cm diameter on bottom, and 30 cm in height) and an inner ring (10 cm  
90 diameter and 10 cm in height) in the middle. Some small holes were drilled on the  
91 bottom of inner ring for drainage. The outer ring and the inner ring were connected  
92 to a smooth slope. An outlet was installed at the end of slope for water and sediment  
93 collection. The detailed description of rainfall device can be found in Xiao et al.  
94 (2017, 2018).

95 Before packing soil in the inner ring, the ring bottom was filled with some gravel  
96 with diameters of 1-2 cm to ensure free drainage. Then a filter paper was covered  
97 over the gravel and air-dried soil was packed over the paper. The packed soil was  
98 about 1.5 cm thick with bulk density of  $1.20 \text{ g cm}^{-3}$  and initial soil water content of  
99 5%. The duration of simulated rainfall was 10 mins with an intensity of  $60 \text{ mm h}^{-1}$ .  
100 The kinetic energy was simulated at four different rainfall heights, 0.5 m, 1.0 m, 1.5  
101 m and 2.0 m. Each treatment replicated twice. The splashed soil sticking on the wall  
102 of outer ring and slope was washed out with an injector, and then collected at the  
103 outlet. The collected sediment was dried and weighted. After test, the remained

104topsoil of 0.5 cm thick in the splash pan was collected to measure the water stable  
105aggregates after air-dried.

106 Non-uniform swelling of soil minerals after wetting has limited damage to  
107aggregates in rainfall conditions (Almajmaie, Hardie, Acuna, & Birch, 2017; Le  
108Bissonnais, 1996). At the same time, the loessal parent material of the test soils was  
109the least swelling mineral. Therefore, the destruction of soil aggregates by non-  
110uniform swelling after soil mineral wetting was neglected in short duration of this  
111study. The aggregates disintegration was mainly caused by slaking and mechanical  
112striking (Xiao et al., 2017; 2018). The aggregates disintegration by 95% ethanol was  
113resulted mainly from the mechanical striking because of the limited slaking effects  
114of ethanol. Two raindrop materials (deionized water and ethanol), were used to  
115distinguish the slaking and mechanical striking.

### 1162.3 Measurement

117 Soil particle size distribution, pH value, and soil organic carbon were analyzed  
118by a pipette method (Liu, 1996), the Rex Electric Chemical PHS-3E precision  
119acidity meter (Shanghai Precision Scientific Instrument Co., Ltd, China) and the  
120potassium dichromate oxidation-external heating method, respectively (Liu, 1996).  
121The content of  $\text{CaCO}_3$  was determined by using a gas volume method (Dreimanis,  
1221962; Zhao et al., 2016). The macroaggregate ( $>0.25$  mm) measurement adopted the  
123wet sieving method improved by Yoder (1936). The light and heavy fractions of  
124organic carbon were separated in NaI solution with the density of  $1.7 \text{ g cm}^{-3}$  (Elliott,  
1251991).

126 LB method was used to measure the aggregate stability under three treatments:  
127fast wetting (FW), slow wetting (SW) and mechanical breakdown by slaking after  
128pre-wetting (WS) (Le Bissonnais, 1996). The soil samples were air dried and 3-5  
129mm aggregates were selected. The 3-5mm aggregates are dried in a  $40 \text{ }^\circ\text{C}$  oven for

13024 hours to ensure they are at the same matrix potential. Then the aggregates were  
 131treated in three different treatments. FW: the aggregates of 5 g were immersed in 50  
 132mL deionized water, and the water was absorbed by pipette after 10 minutes. SW:  
 133the aggregates with 5 g were gently placed on the matrix potential of -0.3 kPa for 30  
 134minutes to ensure that the aggregates were wetted completely. SW: the aggregates of  
 1355 g were immersed in 50 mL ethanol (95% in mass), and the ethanol was absorbed  
 136by pipette after 10 minutes. And then transferred the aggregates to a 250 mL flask  
 137filled with 200 cm<sup>3</sup> deionized water, corked and stirred up and down for 20 times,  
 138and the water was absorbed by pipette after 30 minutes. Transfer the soil aggregates  
 139from the above three treatments to a sieve (0.05 mm) already immersed in alcohol  
 140(95% in mass) and shake up and down 20 times. The aggregate retained in the sieve  
 141was baked for 48 hours in the oven at 40 °C. The dried aggregate was passed through  
 142the dry sieve of 3, 2, 1, 0.5, 0.25, 0.1 and 0.05 mm, and then measured for their size.  
 143Each treatment is repeated three times.

144 Aggregate stability is expressed in terms of mean weight diameter (MWD).

$$145 \quad \text{MWD} = \sum_{i=1}^n w_i x_i \quad (1)$$

146where  $w_i$  is the weight fraction of aggregates in size class  $i$  with an average diameter  
 147 $x_i$ .

#### 1482.4 Data analysis

149 The relative slaking index (RSI) and the relative mechanical breakdown index  
 150(RMI) were used to evaluate the sensitivity of aggregates to slaking and mechanical  
 151breakdown effects (Zhang & Horn, 2001).

$$152 \quad \text{RSI} = \frac{\text{MWD}_{\text{SW}} - \text{MWD}_{\text{FW}}}{\text{MWD}_{\text{SW}}} \quad (2)$$

$$153 \quad \text{RMI} = \frac{\text{MWD}_{\text{SW}} - \text{MWD}_{\text{WS}}}{\text{MWD}_{\text{SW}}} \quad (3)$$

154 where  $\text{MWD}_{\text{FW}}$ ,  $\text{MWD}_{\text{WS}}$ , and  $\text{MWD}_{\text{SW}}$  are the mean weight diameter obtained by  
 155 the FW, WS, and SW treatments, respectively (Le Bissonnais, 1996). The larger of  
 156 RSI or RMI, the higher sensitivity of the aggregates to slaking or mechanical  
 157 breakdown.

158 The splash erosion rate was the splashed-out soil mass from the test area per unit  
 159 area per unit time, which can be calculated with Eq. (4):

$$160 \quad D = \frac{S}{At} \quad (4)$$

161 where  $D$  is the splash erosion rate ( $\text{g m}^{-2} \text{min}^{-1}$ ),  $S$  is the mass of the splashed material  
 162 (g),  $A$  is the test area ( $\text{m}^2$ ) and  $t$  is the duration of the rain (min).

163 The rainfall kinetic energy was calculated by referring to the formula in Xiao et  
 164 al. (2017; 2018). Alcohol and deionized water have different rainfall kinetic energy  
 165 due to their characteristics. The raindrop parameters and rainfall kinetic energy are  
 166 shown in [Table 2](#).

167 All statistical analyses were performed by using Excel 2010 and SPSS 19.0. Soil  
 168 aggregate stability indexes were analyzed with a variance analysis (ANOVA), and  
 169 the others with the Pearson correlation analysis (i.e., splash erosion rate, contribution  
 170 of slaking and mechanical striking, etc.).

## 1713. Results

### 1723.1 Aggregate stability indexes

173 The aggregate stability indexes for the five soils are shown in [Figure 2](#). The  
 174  $\text{MWD}_{\text{FW}}$ ,  $\text{MWD}_{\text{WS}}$  and  $\text{MWD}_{\text{SW}}$  ranged from 0.612 to 2.389, from 1.202 to 3.262,  
 175 and from 1.935 to 3.367, respectively. The MWD values increased with the increase

176of SOC contents. The aggregate stability values increased in the order of  $MWD_{SW}$   
 177 $>MWD_{WS} >MWD_{FW}$  for the five soils. It showed that the effect of chemical  
 178dispersion (SW) was the weakest of aggregate breakdown mechanisms, whereas  
 179slaking (FW) had the most effect on aggregate breakdown. The RSI and RMI  
 180decreased from 0.698 to 0.293, and from 0.325 to 0.033 with the increase of SOC,  
 181respectively. The value of RSI was larger than that of RMI for the five soils.

182 [Figure 3](#) indicated that  $MWD_{FW}$ ,  $MWD_{WS}$  and  $MWD_{SW}$  had significant positive  
 183correlations with SOC contents, but no significant positive correlation with LFOC  
 184and HFOC. They were negatively correlated with clay and the content of  $CaCO_3$ .  
 185 $MWD_{FW}$ ,  $MWD_{WS}$  and  $MWD_{SW}$  had no significant positive relationships with the  
 186contents of free-form Fe, amorphous Fe, free-form Al and amorphous Al. RSI and  
 187RMI had significant negative correlations with the contents of SOC and LFOC,  
 188while they had no significant corrections with free-form Fe, amorphous Fe, free-  
 189form Al and amorphous Al contents.

### 1903.2 Splash erosion rate

191 Splash erosion rate increased with the increase of rainfall kinetic energy for both  
 192deionized water and alcohol raindrops ([Figure 4](#)). The relationships for five soils  
 193could be described by power functions, and the coefficient of determination ( $R^2$ ) was  
 194higher than 0.94 for both deionized water and ethanol tests ([Table 3](#)). The coefficient  
 195of power function can serve as an indicator of erosion severity with higher values  
 196reflecting higher soil erodibility.

197 The splash erosion rate has no significant negative correlations with SOC and  
 198HFOC in both deionized water and ethanol tests ([Table 4](#)). The negative and positive  
 199correlations were found between splash erosion rate and LFOC in deionized water  
 200and ethanol tests, respectively. An exception was that the negative correlation was  
 201found for ethanol tests in 1.5 m rainfall height. Meanwhile, compared with SOC and

202HFOC, LFOC had weaker relations with splash erosion rate in deionized water tests.

### 2033.3 Macroaggregates

204 The macroaggregate (>0.25mm) contents remained in splash pan after rainfall in  
 205 ethanol tests were more than that in deionized water tests (Figure 5). The  
 206 macroaggregate contents for eroded soils were less than those of the parent soil for  
 207 both deionized water and ethanol tests. However, the macroaggregate contents were  
 208 increasingly closer to the parent soil with the increase of soil organic carbon  
 209 contents, and the trend was more obvious in ethanol tests. The macroaggregate  
 210 contents decreased with the increase of kinetic energy in both deionized water and  
 211 ethanol tests, whereas the kinetic energy had no such significant effects in the  
 212 ethanol tests except for soil sample IV.

213 The positive correlations were found between the macroaggregate contents and  
 214 SOC, LFOC and HFOC in deionized water tests (Table 4). The correlations between  
 215 macroaggregate and HFOC were weaker than those of SOC and LFOC. However,  
 216 there were no significant correlations between them in ethanol tests.

### 2173.4 Effects of slaking and mechanical striking on splash erosion

218 Figure 6 showed that the contribution rate of slaking and mechanical striking  
 219 decreased from 75% to 25% and increased from 25% to 75% with the increase of  
 220 rainfall kinetic energy, respectively. Meanwhile, when the rainfall kinetic energy was  
 221 less than the range of critical values (between  $9 \text{ J m}^{-2} \text{ mm}^{-1}$  and  $12 \text{ J m}^{-2} \text{ mm}^{-1}$ ), the  
 222 contribution of slaking has dominant impact on aggregates disintegration. When the  
 223 rainfall kinetic energy was greater than the range of critical values, the contribution  
 224 of mechanical striking to splash erosion is gradually greater than that of slaking.

225 Table 4 indicated that the contribution rate of slaking had negative correlations  
 226 with SOC contents when the kinetic energy increased from 3 to  $12 \text{ J m}^{-2} \text{ mm}^{-1}$ , and it  
 227 had positive correlations when the kinetic energy changed between 15 to  $18 \text{ J m}^{-2}$

228mm<sup>-1</sup>. With the increase of the kinetic energy, the correlation coefficient decreased  
 229from -0.774 to 0.061. There were no significant correlations between the  
 230contribution rate of slaking and LFOC, meanwhile the correlation coefficient had an  
 231increasing trend with the increase of the kinetic energy. The correlations between  
 232contribution rate of slaking and HFOC was significantly negative when the kinetic  
 233energy increased from 3 to 6 J m<sup>-2</sup> mm<sup>-1</sup>. The correlation coefficient decreased from  
 2340.900 to 0.671 with the increase of kinetic energy.

#### 2354. Discussion

236 Generally, soil clay, SOC, CaCO<sub>3</sub>, and Fe/Al oxides act as cementing agents  
 237that affect the formation and stability of aggregates (An, Darboux, & Cheng, 2013;  
 238Dimoyiannis, 2012; Le Bissonnais., 1996; Le Bissonnais & Arrouays, 1997). The  
 239aggregate stability had significant positive correlation with SOC but not with the  
 240contents of clay, CaCO<sub>3</sub>, and Fe/Al oxides (Figure 3). In this study, SOC acted as the  
 241main factor affecting the aggregate stability because the test soils had the similar  
 242contents of clay, CaCO<sub>3</sub>, and Fe/Al oxides (Table 1). The SOC and LFOC had  
 243significantly negative correlations with RSI and RMI, illustrating that the sensitivity  
 244of slaking and mechanical striking decreased with increases of SOC and LFOC.  
 245Therefore, SOC, especially the LFOC, played an important role in resisting  
 246disintegration of aggregates. The formation of soil aggregate relies on organic  
 247materials, and the organic binding agents were mainly polysaccharides, roots and  
 248fungal hyphae, strongly sorbed natural polymers, and so on (Sdall & Oades, 1982).  
 249The roots and fungal in composition of LFOC could promote the formation of soil  
 250aggregate directly (Elliott, 1986; Oades, 1984). Thus, the aggregate stability and the  
 251organic binding agents increased with the content of LFOC.

252 The power function relationships between splash erosion rate and rainfall kinetic  
 253energy is consistent with the conclusions of the previous researchers (Hu, Zhen, &

254Bian, 2016; Sharma, Gupta, & Rawls, 1991; Xiao et al., 2017; 2018). There was a  
255tendency that splash erosion rate was negatively correlated with SOC, LFOC and  
256HFOC for both deionized water and ethanol tests. However, the correlations were  
257not statistically significant. This could be caused by the relatively narrow ranges of  
258SOC, LFOC and HFOC used in this study, or splash erosion might be not as  
259sensitive to SOC and LFOC as the aggregate sensitivity to slaking and mechanical  
260breakdown effects.

261 Raindrops hit the soil surface with a certain kinetic energy, which is often  
262sufficient to breakdown soil aggregates and compact the soil surface (Moss & Green,  
2631987). The deionized water raindrops had both slaking and mechanical striking  
264effects on aggregate disintegration, whereas alcohol only had mechanical striking  
265effects (Le Bissonnais, 1996). On the other hand, the kinetic energy of deionized  
266water raindrops was greater than that of ethanol raindrops at the same fall height  
267([Table 2](#)). These results lead to the destructive capacity of deionized water raindrops  
268were greater than that of ethanol.

269 The mechanical striking of raindrops on soil could be greatly reduced by  
270vegetation cover (Lal, 1976; Adekalu, Olorunfemi, & Osunbitan, 2007; Kukul &  
271Sarkar, 2010). However, vegetation could promote the accumulation of SOC,  
272especially the LFOC in short term (Boone, 1994; Garcia, Hemanderz, Roldan, &  
273Martin, 2002; Gil-Sotres, Trasar-Cepeda, Leiros, & Seoane, 2005). That resulted in  
274the increase of aggregate stability ([Figure 2](#)), and counteracted the increase of  
275slaking contribution. Finally, vegetation coverage could improve soil antierodibility  
276by reducing both slaking and mechanical striking effects of raindrops.

277 There are some limitations for testing five soils developed from only one parent.  
278The effects of SOC on soil aggregates disintegration may be different for different  
279soil types due to interactive effects of other factors. Furthermore, the coupled effects

280of other factors in aggregate breakdown during splash erosion also need to be  
281researched in the future.

## 2825. Conclusions

283 In this study, the simulated rainfall experiments for five soils with different SOC  
284were carried out. The results indicated that the content of SOC and LFOC had  
285substantial effects on aggregate stability. The RSI and RMI decreased as SOC and  
286LFOC increased. The amount of macroaggregates in deionized water tests were  
287significantly less than that in alcohol tests under the same rainfall heights. The  
288reduction of macroaggregates in eroded soil gradually decreased with the increase of  
289SOC and LFOC, especially in alcohol test. As the rainfall kinetic energy increased,  
290the contribution of slaking to soil splash decreased while the contribution of  
291mechanical striking increased. The range of critical values between  $9 \text{ J m}^{-2} \text{ mm}^{-1}$  and  
292 $12 \text{ J m}^{-2} \text{ mm}^{-1}$  were found to determine the dominated contribution of slaking and  
293mechanical striking to splash erosion.

294**Data sharing:** Research data are not shared.

295**Conflict of interest:** none

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## 472 Tables

473

Tab.1 Basic physical-chemical properties of five test soils in the experiments

Soil samples	Year of vegetation restoration / a	Longitude and latitude	Clay / %	Silt / %	Sand / %	CaCO <sub>3</sub> / g kg <sup>-1</sup>	Soil moisture content/ %	Soil organic carbon/ g kg <sup>-1</sup>	Light fractions of soil organic carbon/ g kg <sup>-1</sup>	Heavy fractions of soil organic carbon/ g kg <sup>-1</sup>	pH (1:2.5)	Free-form Fe/g kg <sup>-1</sup>	Amorphous Fe/g kg <sup>-1</sup>	Free-form Al/g kg <sup>-1</sup>	Amorphous Al/g kg <sup>-1</sup>
I	3	N36°03.61', E109°08.99'	22.36	21.16	56.48	81.09	2.46	7.54	5.89	1.68	8.51	5.10	0.48	1.57	0.58
II	20	N36°03.60', E109°09.03'	23.54	25.23	51.23	82.53	1.18	13.34	10.45	2.86	8.36	7.91	0.45	0.73	1.17
III	100	N36°04.94', E109°08.71'	20.18	29.63	50.20	77.63	0.72	14.85	9.83	4.99	8.42	8.93	0.64	0.56	1.31
IV	80	N36°03.68', E109°08.87'	21.05	18.06	60.88	76.76	1.30	18.39	15.70	2.66	8.25	6.16	0.81	1.87	1.63
V	55	N36°03.85', E109°08.78'	22.12	15.84	62.04	79.98	0.92	21.69	17.88	3.78	8.37	4.66	0.39	1.70	0.44

Tab. 2 Rainfall kinetic energy for different fall heights

Liquid	Fall height/m	Time for 10 raindrops/s	Weight of 10 raindrops/g	Mean raindrop diameter/mm	Rainfall kinetic energy/J m <sup>-2</sup> mm <sup>-1</sup>
Deionized water	0.5	9.91	0.09	2.62	1.77
	1.0	9.91	0.09	2.63	6.15
	1.5	9.91	0.09	2.62	12.06
	2.0	9.92	0.09	2.63	18.75
Ethanol	0.5	6.45	0.04	2.03	1.49
	1.0	6.23	0.04	2.03	5.00
	1.5	6.34	0.04	2.03	8.90
	2.0	6.29	0.04	2.03	13.06

475 Tab. 3 Nonlinear regression between rainfall kinetic energy and splash erosion rate in simulated rainfall tests by using deionized water and  
 476 ethanol as raindrop

Number of soil samples	Deionized water raindrop	R <sup>2</sup>	Ethanol raindrop	R <sup>2</sup>
I	$D = 1.1959 e^{1.15}$	0.97	$D = 0.3055 e^{1.38}$	0.98
II	$D = 2.2951 e^{0.89}$	0.99	$D = 0.5391 e^{1.27}$	0.99
III	$D = 0.7207 e^{1.18}$	0.99	$D = 0.2026 e^{1.51}$	0.98
IV	$D = 1.8037 e^{1.03}$	0.97	$D = 0.5753 e^{1.19}$	0.94
V	$D = 0.6303 e^{1.26}$	0.99	$D = 0.2111 e^{1.41}$	0.96

Notes: D is the soil splash rate ( $\text{g m}^{-2} \text{min}^{-1}$ ) and e is the rainfall kinetic energy ( $\text{J m}^{-2} \text{mm}^{-1}$ )

477 Tab. 4 Pearson correlation coefficients between soil organic carbon and its fractions related to the splash erosion rate, contribution of slaking and  
478 macroaggregates (>0.25mm) contents

	Splash erosion rate								Contribution of slaking						Macroaggregates (>0.25mm) contents							
	Deionized water				Ethanol				C <sub>3</sub>	C <sub>6</sub>	C <sub>9</sub>	C <sub>12</sub>	C <sub>15</sub>	C <sub>18</sub>	Deionized water				Ethanol			
	D <sub>0.5</sub>	D <sub>1.0</sub>	D <sub>1.5</sub>	D <sub>2.0</sub>	D <sub>0.5</sub>	D <sub>1.0</sub>	D <sub>1.5</sub>	D <sub>2.0</sub>							W <sub>0.5</sub>	W <sub>1.0</sub>	W <sub>1.5</sub>	W <sub>2.0</sub>	W <sub>0.5</sub>	W <sub>1.0</sub>	W <sub>1.5</sub>	W <sub>2.0</sub>
SOC	-0.251	-0.270	-0.310	-0.513	-0.036	-0.031	-0.502	-0.026	-0.774	-0.334	-0.126	-0.005	0.016	0.061	0.632	0.429	0.616	0.656	0.083	0.036	-0.144	-0.016
LFOC	-0.158	-0.086	-0.131	-0.363	0.095	0.056	-0.490	0.120	-0.623	-0.129	0.072	0.188	0.201	0.245	0.616	0.431	0.606	0.619	0.177	-0.129	-0.072	0.050
HFOC	-0.468	-0.822	-0.818	-0.791	-0.525	-0.353	-0.263	-0.576	-0.900*	-0.918*	-0.802	-0.733	-0.692	-0.671	0.316	0.163	0.289	0.408	-0.328	-0.349	-0.343	-0.267

Note: D<sub>0.5</sub>, D<sub>1.0</sub>, D<sub>1.5</sub>, D<sub>2.0</sub> is splash erosion rate at different fall heights (0.5m, 1.0m, 1.5m and 2.0m), respectively; C<sub>3</sub>, C<sub>6</sub>, C<sub>9</sub>, C<sub>12</sub>, C<sub>15</sub> and C<sub>18</sub> is the contribution of slaking in different rainfall kinetic energy (3, 6, 9, 12, 15 and 18 J m<sup>-2</sup> mm<sup>-1</sup>), respectively. SOC, LFOC and HFOC is soil organic carbon, light fractions organic carbon and heavy fractions organic carbon content, respectively. W<sub>0.5</sub>, W<sub>1.0</sub>, W<sub>1.5</sub> and W<sub>2.0</sub> is the contents of >0.25mm water stable aggregates at different heights (0.5m, 1.0m, 1.5m and 2.0m).

\* Significant at 0.05 level of probability.

479 **Figure captions**

480 Figure 1. Schematic representation of the experiment device

481 Figure 2. Aggregates water-stability of five loessial soils developed from same parent

482 material with different soil organic carbon (SOC) contents. The SOC contents of

483 I, II, III, IV, and V is 7.54, 13.34, 14.85, 18.39, 21.69 g kg<sup>-1</sup>, respectively

484 Different letters in the same set of data of the same color indicate significant

485 differences at 5% level.

486 Figure 3. Heatmap for the relationships between soil aggregate stability indexes and

487 soil properties.  $MW_{SD}$  and  $MW_{SW}$  denote the mean weight

488 diameters obtained after the fast-wetting (FW), pre-wetting and stirring (WS)

489 and slow wetting (SW), respectively; RSI and RMI denote relative slaking index

490 and relative mechanical breakdown index, respectively; SOC, LFOC and HFOC

491 denote soil organic carbon, light fractions organic carbon and heavy fractions

492 organic carbon content, respectively.

493 Figure 4. Relationships between rainfall kinetic energy of two kind of raindrops (A is

494 deionized water; B is ethanol) and splash erosion rate. I, II, III, IV and V were

495 five different tested soils, which were developed from the same

496 similar physicochemical properties and different organic carbon content because

497 of different vegetation restoration time.

498 Figure 5. The contents of macroaggregates (>0.25mm) in parent soil and

499 remained in splash pan after rainfall at different height (0.5, 1.0, 1.5, and 2.0 m)

500 with different raindrops (A is deionized water; B is ethanol). Different letters in

501 the same group of each soil sample indicate significant differences at the 0.05

502 level.

503 Figure 6. Changes of contribution rate of slaking and mechanical striking (S is the  
504 contribution rate of slaking; M is the contribution rate of mechanical striking) to  
505 splash erosion with rainfall kinetic energy. I, II, III, IV and V were five different  
506 tested soils, which were developed from the s  
507 physicochemical properties and different organic carbon conten  
508 different vegetation restoration time.